1	High rejection reverse osmosis membrane for
2	removal of <i>N</i> -nitrosamines and their precursors
3	Revised manuscript submitted to
4	Water Research
5	December 2017
6	Takahiro Fujioka <sup>1,*</sup> , Kenneth P. Ishida <sup>2</sup> , Takuji Shintani <sup>3</sup> , Hitoshi Kodamatani <sup>4</sup>
7	<sup>1</sup> Water and Environmental Engineering, Graduate School of Engineering,
8	Nagasaki University, 1-14 Bunkyo-machi, Nagasaki 852-8521, Japan
9	<sup>2</sup> Research & Development Department, Orange County Water District, 18700 Ward Street,
10	Fountain Valley, CA 92708, USA
11	<sup>3</sup> Division of Advanced Membrane Science and Technology, Graduate School of Science,
12	Technology and Innovation, Kobe University, 1-1 Rokkodai-cho, Kobe 657-8501, Japan
13	<sup>4</sup> Division of Earth and Environmental Science, Graduate School of Science and Engineering,
14	Kagoshima University, 1-21-35 Korimoto, Kagoshima 890-0065, Japan

<sup>16 \*</sup> Corresponding author: Takahiro Fujioka, Email: tfujioka@nagasaki-u.ac.jp, Ph +81 095 819 2695

#### 17 Abstract

18 Direct potable reuse is becoming a feasible option to cope with water shortages. It requires 19 more stringent water quality assurance than indirect potable reuse. Thus, the development of 20 a high-rejection reverse osmosis (RO) membrane for the removal of one of the most 21 challenging chemicals in potable reuse – N-nitrosodimethylamine (NDMA) – ensures further 22 system confidence in reclaimed water quality. This study aimed to achieve over 90% removal 23 of NDMA by modifying three commercial and one prototype RO membrane using heat 24 treatment. Application of heat treatment to a prototype membrane resulted in a record high removal of 92% (1.1-log) of NDMA. Heat treatment reduced conductivity rejection and 25 26 permeability, while secondary amines, selected as N-nitrosamine precursors, were still well 27 rejected (>98%) regardless of RO membrane type. This study also demonstrated the highly 28 stable separation performance of the heat-treated prototype membrane under conditions of 29 varying feed temperature and permeate flux. Fouling propensity of the prototype membrane 30 was lower than a commercial RO membrane. This study identified a need to develop highly 31 selective RO membranes with high permeability to ensure the feasibility of using these 32 membranes at full scale.

33 Keywords: *N*-nitrosodimethylamine; *N*-nitrosamine precursor; heat treatment; reverse
34 osmosis; potable water reuse.

## 36 1. Introduction

37 Potable reuse (PR) has been increasingly considered as a viable and powerful option to cope 38 with water scarcity in many parts of the world (Burgess et al., 2015). Most PR schemes 39 employed over the past decade are based on indirect potable reuse (IPR), which is 40 implemented through the augmentation of drinking water sources (e.g. dams and aquifers) 41 with highly treated wastewater. To make PR more economically feasible, the water industry 42 has devoted significant attention on direct potable reuse (DPR) (Arnold et al., 2012). In DPR, 43 recycled water with a short retention time is transported directly to the entrance of drinking 44 water treatment plants, so that capital and operating costs associated with infrastructure and 45 water quality monitoring can be reduced considerably. One of the critical considerations 46 when shifting from IPR to DPR is the quality assurance of recycled water (Drewes and Khan, 47 2015; Leverenz et al., 2011). In fact, a feasibility study of DPR initiated by the California 48 State Legislature (USA) has recommended greater focus on the identification and removal of 49 low molecular weight trace organic compounds (TOrCs) (CSWRCB, 2016).

Among low molecular weight TOrCs, N-nitrosodimethylamine (NDMA; C<sub>2</sub>H<sub>6</sub>N<sub>2</sub>O) - a 50 51 probable carcinogenic chemical (USEPA, 1993) – is one of the more challenging compounds 52 in PR. For example, an adsorption process using granular activated carbon only removed NDMA up to 50% (Fleming et al., 1996; Schmidt and Brauch, 2008). To date, NDMA is 53 54 essentially removed by a combination of direct photolysis by UV irradiation and an advanced 55 oxidation process (AOP) that is based on hydrogen peroxide dosage (Leverenz et al., 2011). 56 California has established a notification level (NL) of 10 ng/L for NDMA and a public health 57 goal (PHG) of 3 ng/L. No reliable removal credits for NDMA have been granted for the 58 application of a reverse osmosis (RO) process.

59 NDMA, an N-nitrosamine, is a hydrophilic compound with molecular weight of 74 g/mol and 60 is uncharged at pH 6-8. Because of its small molecular size and uncharged property, NDMA 61 readily permeates through an RO membrane. Brackish water RO membranes are typically 62 able to achieve 40-70% rejection of NDMA at lab-scale levels (Bellona et al., 2011; Fujioka 63 et al., 2012b; Hofs et al., 2013; Miyashita et al., 2009; Steinle-Darling et al., 2007) and at 5-64 80% at pilot- to full-scale levels (Bellona et al., 2008; Farré et al., 2011a; Fujioka et al., 2013b; Plumlee et al., 2008; Poussade et al., 2009). Any improvement to the selectivity of 65 66 RO membranes and assignment of removal credits for NDMA can improve the safety of 67 recycled water and possibly reduce the load on any UV or UV/AOP.

68 Formation of NDMA downstream of the RO process has also gained increased concern in 69 DPR. NDMA can form through a reaction between residual chloramine and NDMA 70 precursors in RO permeate (McCurry et al., 2017; Sgroi et al., 2015; Soltermann et al., 2013). 71 These NDMA precursors include dimethylamine (DMA) and tertiary amines (Mitch et al., 72 2003; Schreiber and Mitch, 2005; Shah and Mitch, 2011). Although these precursors are well 73 removed by RO membranes (e.g. >99%) (Krauss et al., 2010; Miyashita et al., 2009), 74 significant NDMA precursors still remain in RO permeate (Farré et al., 2011b) that are not degraded by UV/AOP (or are reaction products). Thus, the development of a highly selective 75 76 RO membrane for enhanced removal of NDMA and NDMA precursors could help to 77 improve on the safety of recycled water in DPR. In addition, improvement of the selectivity 78 of RO membranes is expected to become more important in PR to mitigate future water 79 quality issues associated with non-regulated or unidentified emerging TOrCs (Debroux et al., 80 2012; Werber et al., 2016a).

To improve the rejection of RO membranes, a simple membrane modification technique based on an immersion of RO membranes in high-temperature ultrapure water, has been proposed (Fujioka et al., 2015). Heat treatment can enhance the rejection of uncharged and low molecular weight compounds (e.g. boric acid); however, water permeability also decreases. Heat treatment during the interfacial polymerisation process has the effect of tightening the membrane structure and improving its salt rejection (Shintani et al., 2009). To the best of our knowledge, there are still no RO membranes with a reported ability to remove NDMA >90% (1.0-log).

This study aimed to achieve over 1.0-log removal of NDMA by modifying three commercial and one prototype RO membrane using heat treatment. In addition to NDMA, this study included five additional *N*-nitrosamines and five secondary amines to demonstrate the separation performance and stability of heat-treated RO membranes for a range of feed temperatures and permeate flux. Fouling propensity of heat-treated RO membranes was also examined using treated wastewater. Lastly, the feasibility and implication of using high rejection RO membranes for full-scale DPR schemes are discussed.

## 96 2. Materials and methods

#### 97 2.1. Chemicals

98 Certified 100 mg/L solutions of six analytical grade N-nitrosamines (Table S1) - NDMA, N-99 nitrosomethylethylamine (NMEA), N-nitrosopyrrolidine (NPYR), N-nitrosodiethylamine 100 (NDEA), N-nitrosopiperidine (NPIP) and N-nitrosomorpholine (NMOR) – were purchased 101 from Ultra Scientific (Kingstown, RI, USA). Five analytical grade secondary amines – DMA 102 hydrochloride, pyrrolidine (PYR), diethylamine (DEA), piperidine (PIP) and morpholine (MOR) - selected as N-nitrosamine precursors were purchased from Tokyo Chemical 103 104 Industry (Tokyo, Japan). Physicochemical properties of these N-nitrosamines and secondary 105 amines are summarised in Table 1. The N-nitrosamines and secondary amines, all of which 106 have a Log D value of <2.0, are classified as hydrophilic compounds (Van der Bruggen et al.,

107 2006). pK<sub>a</sub> values of N-nitrosamines are well below 8; thus, they are uncharged at pH 8. In 108 contrast, the secondary amines (except MOR) are all dissociated at pH 8 with pKa values well 109 beyond 8. MOR is dissociated by ~75% at pH 8. Among the selected N-nitrosamines, NDMA 110 has the lowest minimum projection area, i.e., the area of the compound projected with the minimum plane of its circular disk (Fig. S2). Stock solutions containing four or six 111 112 *N*-nitrosamines in methanol were prepared at 1 µg/mL of each compound. Stock solutions of each secondary amine were prepared at 100 mM in ultrapure water. All stock solutions were 113 114 stored at 4 °C in the dark. Analytical grade NaCl, CaCl<sub>2</sub> and NaHCO<sub>3</sub> were purchased from 115 Wako Pure Chemical Industries (Tokyo, Japan). Activated sludge effluent from a municipal 116 wastewater treatment plant in Japan was treated by ultrafiltration (UF) and used for fouling 117 experiments. Total organic carbon, pH and conductivity of the UF-treated wastewater was 118 6.5 mg/L,  $6.5 \text{ and } 1112 \mu\text{S/cm}$ , respectively.

119

#### [Table 1]

#### 120 2.2. Membranes and membrane treatment system

121 Three commercial brackish water RO membranes – ESPA2, ESPAB and HYDRApro<sup>®</sup>501 – 122 and one prototype membrane were supplied as flat sheet samples by Hydranautics/Nitto 123 (Osaka, Japan). All of the RO membranes are thin-film composite polyamide. ESPA2 is 124 commonly used in water recycling applications (Fujioka et al., 2012a). ESPAB is employed 125 in the second stage of seawater desalination to achieve a high rejection of boron. 126 HYDRApro<sup>®</sup>501 (HYDRA) is an RO membrane that is designed for industrial process 127 applications. The prototype membrane (Prototype) is a proprietary RO membrane.

128 A bench-scale RO treatment system was comprised of a stainless steel membrane cell (Iwai 129 Pharma Tech, Tokyo, Japan), high-pressure pump (KP-12, FLOM, Tokyo, Japan), 2-L glass 130 reservoir with a stainless steel heat exchanging coil connected to a temperature control unit 131 (NCB-500, Tokyo Rikakikai, Tokyo, Japan) (Fig. S3). The membrane cell held a circular
132 flat-sheet membrane coupon with effective surface area of 36.3 cm<sup>2</sup>.

## 133 2.3. Experimental protocols

134 2.3.1. Heat treatment

The RO membrane coupons were placed in 200 mL beakers and rinsed with copious amounts of ultrapure water (18.0 M $\Omega$ cm). Membranes were heat-treated by placing the beakers in a temperature-controlled water bath (SWB-11A, AS ONE, Osaka, Japan) at 80 ± 1 °C for 4 h. After heat treatment, the heat-treated RO membranes were rinsed with deionized water and stored at 4 °C in the dark.

## 140 2.3.2. Separation evaluation

The performance of untreated and heat-treated RO membranes was evaluated in clean water. 141 142 RO membranes were first conditioned and stabilised with a deionized water at 1500 kPa. 143 Stock solutions of background electrolytes were dosed at 20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 144 1 mM CaCl<sub>2</sub> in the RO feed. N-nitrosamines were added to the RO feed at a concentration of 1000 ng/L. N-nitrosamines were evaluated separately from the secondary amines. 145 146 Concentrations of secondary amines in RO feed were adjusted at 50 µg/L. After conditioning, the RO system was operated at 20 L/m<sup>2</sup>h permeate flux, which is typically employed in water 147 148 recycling applications (Fujioka et al., 2012a). The feed temperature was 20 °C. Following 1 h of operation, RO feed and permeate samples were collected in 1.5 mL amber vials for the 149 150 analysis of N-nitrosamines and in 10 mL vials for the analysis of secondary amines. The N-151 nitrosamines and secondary amines used in this study are very hydrophilic at pH 8; thus, 152 adsorption of these chemicals onto the membrane was unlikely to occur. In fact, previous 153 studies reported that the rejection of N-nitrosamines by RO membranes in clean water or 154 wastewater matrices reached a steady state condition within 1 h of filtration (Fujioka et al.,

155 2013a; Miyashita et al., 2009; Steinle-Darling et al., 2007). Thus, 1 h of filtration was156 assumed to be sufficient to reach adsorption equilibrium.

157 2.3.3. Fouling development

Fouling propensity of each RO membrane was evaluated by operating the system on a secondary wastewater effluent. After conditioning the RO membrane, the deionized water was replaced with UF-treated wastewater. Fouling development was accelerated by operating the system at a high constant flux of 40 L/m<sup>2</sup>h with the feed temperature maintained at 20 °C. Transmembrane pressure (TMP) was recorded.

## 163 2.4. Analytical techniques

#### 164 2.4.1. *N*-nitrosamines

165 determined Concentrations of *N*-nitrosamines were by high-performance liquid 166 chromatography-photochemical reaction-chemiluminescence (HPLC-PR-CL). Details of this 167 technique are provided elsewhere (Fujioka et al., 2016; Kodamatani et al., 2016). The 168 analysis was performed with an eluent of 10 mM phosphate buffer and methanol (90:10 v/v) 169 and the sample injection volume was 200 µL. Clean water matrix samples and permeate 170 samples were analysed without any pretreatment. UF-filtered wastewater collected from the 171 RO feed was prefiltered with a 0.45-µm hydrophilic PTFE syringe filter (Filtstar, Starlab 172 Scientific, China) prior to the analysis.

173 2.4.2. Secondary amines

174 Concentrations of secondary amines were determined by HPLC-PR-CL after nitrosation 175 reaction of samples. Details of this technique are provided elsewhere (Kodamatani et al., 176 2017). The nitrosation reaction was conducted by adding the sample solution (2.0 mL), 177 glacial acetic acid (100  $\mu$ L, 0.8 M), and sodium nitrite (100  $\mu$ L, 200 mM) to a 10-mL screw-178 cap glass tube and then stored at 80 °C for 1 h. The solution was treated with sulfamic acid (400 µL, 1.5 M) and sodium carbonate solution (1.4 mL, 1 M) prior to the analysis by HPLCPR-CL. The sample injection volume was 20 µL.

181 2.4.3. Zeta potential

182 The zeta potential of the RO membrane surface was measured at pH 8 in a 1 mM KCl 183 background electrolyte solution using a SurPASS electrokinetic analyser (Anton Paar GmbH, 184 Graz, Austria). The surface zeta potential was calculated from the measured streaming 185 potential of the membrane using the Fairbrother-Mastin method.

## 186 **3. Results and discussion**

## 187 3.1. NDMA Rejection

Rejection of NDMA by untreated and heat-treated RO membranes was >50% at 20 L/m<sup>2</sup>h permeate flux. Among the untreated RO membranes, the Prototype revealed the highest NDMA rejection (84%), followed by HYDRA (74%), ESPAB (62%) and ESPA2 (53%) (Fig. 1). Although the Prototype membrane had the highest NDMA separation capability, the TMP required to achieve 20 L/m<sup>2</sup>h permeate flux was as high as 1.54 MPa in contrast with 0.49, 0.65 and 1.29 MPa for the ESPA2, ESPAB and HYDRA membranes, respectively (Fig. S4).

195

#### [Fig. 1]

Heat treatment applied to the RO membranes considerably enhanced their separation performance. For example, NDMA rejection of the heat-treated ESPA2 membrane increased from 53 to 62% (**Fig. 1**). Likewise, NDMA rejection by the heat-treated ESPAB and HYDRA membranes increased by from 62 to 79% and from 74 to 88%, respectively. Finally, the heat-treated Prototype membrane rejected 92% of the NDMA, achieving the primary objective of this study (i.e. >90%). Enhanced performance from heat treatment was also

observed for the *N*-nitrosamines: NMEA, NPYR and NMOR (**Fig. S4**). Nevertheless, the TMP required to achieve a 20 L/m<sup>2</sup>h permeate flux increased considerably—17% for ESPA2 and >40% for the other membranes (**Fig. S4**). The results indicate that heat treatment applied to RO membranes can reduce the permeation of *N*-nitrosamines at the expense of membrane permeability. The heat treatment of the RO membranes is believed to induce a change in the polymer configuration where upon cooling the end result is a more compact structure of the polyamide separations layer.

209 In contrast to N-nitrosamines, heat treatment resulted in a reduction in conductivity rejection. 210 Heat treatment applied to ESPA2 membrane reduced conductivity rejection from 99.0 to 211 98.8% (Fig. 1). Reduction in conductivity rejection after heat treatment was also apparent for 212 the other three RO membranes (i.e. ESPAB, HYDRA and Prototype). The permeation of ions 213 through an RO membrane is primarily governed by size exclusion and electrostatic repulsion 214 (Bellona et al., 2004). The loss of conductivity rejection by heat treatment may be due to 215 changes in the internal structure. Although a reduction in the negative charge on the 216 membrane surface was also expected to be a contributing factor to reduced conductivity 217 rejection, no definitive conclusions can be made from the charge data (zeta potential) because 218 the negative charge on the membrane surface increased, except for the ESPA2 membrane 219 (Table 2).

220

#### [Table 2]

Heat treatment reduced the pure water permeability of the ESPA2, ESPAB, HYDRA and Prototype RO membranes by 21%, 33%, 35% and 31%, respectively (**Table 2**), indicating that RO membranes could lose 20–35% of their water permeability by heat treatment in exchange for improved rejection. Among eight RO membranes tested in this study, a high correlation ( $R^2 = 0.998$ ) between pure water permeability and NDMA rejection was observed (Fig. 2). The trade-off between permeability and separation performance is often reported in literature (Bernstein et al., 2011; Geise et al., 2011; Werber et al., 2016b; Zhang and Geise, 2016). In contrast, conductivity rejection revealed no correlation with pure water permeability. Providing detailed mechanisms to explain the correlations between *N*-nitrosamine rejection and conductivity rejection requires characterisation of their membrane properties (e.g. free-volume, hole-size and surface charge) (Dražević et al., 2014; Freger, 2015) and will be the scope of future studies.

233

#### [Fig. 2]

## 234 *3.2.* Separation performance

235 The separation capability of the heat-treated Prototype membrane was further evaluated by 236 comparison with the ESPA2 membrane. The ESPA2 membrane was selected as a reference 237 because it is widely used in potable reuse (Fujioka et al., 2012a) and thus, served as a 238 comparison with the best membrane for separations (i.e. heat-treated Prototype). For both RO membranes, rejection of the six N-nitrosamines increased with increasing minimum 239 240 projection area (Fig. 3). This confirmed that size exclusion is the dominant mechanism for 241 the removal of N-nitrosamines. In contrast to the uncharged N-nitrosamines, secondary 242 amines, which are N-nitrosamine precursors, are mostly present in the form of charged spices 243 at pH 8. As a result of electrostatic repulsion interaction in addition to the size exclusion 244 mechanism, their rejection was very high (>98%) for both RO membranes (Fig. 4). DMA, the 245 smallest compound tested, revealed high rejection-98.2% for ESPA2 and 98.9% for the 246 heat-treated Prototype membrane. In addition, heat treatment enhanced the rejection of three 247 secondary amines (MOR, PYR and PIP) to near-complete removal levels (i.e. >99.9%). 248 Overall, the separation capacity of the heat-treated Prototype membrane for secondary amines 249 was greater than the ESPA2 membrane.

250

#### [Fig. 3]

251

## [Fig. 4]

252 A seasonal change in feed temperature could alter the separation performance; thus, stability 253 in the separation performance of the untreated ESPA2 and heat-treated Prototype membrane 254 under conditions of variable feed temperature was evaluated. The variation in N-nitrosamine 255 rejection was significant for the ESPA2 membrane. An increase in feed temperature from 10 256 to 30 °C led to a considerable drop in NDMA rejection from 59 to 39% (Fig. 5). In contrast, 257 far less impact of feed temperature was observed with the heat-treated Prototype membrane 258 as NDMA rejection only dropped from 94 to 90% when the temperature increased from 10 to 30 °C, The rejection of the other N-nitrosamines remained almost constant across the 10-259 260 40 °C range.

261

#### [Fig. 5]

262 The impact of varying the permeate flux on N-nitrosamine rejection was also evaluated. 263 Similar to the case of feed temperature above, the rejection of N-nitrosamines by the heat-264 treated Prototype membrane remained high (Fig. 6). A reduction in permeate flux from 20 to 265 10 L/m<sup>2</sup>h resulted in a decrease in NDMA rejection by ESPA2 membrane from 53 to 42%, 266 while NDMA rejection by the heat-treated Prototype membrane decreased from 92 to 88%. 267 Impacts of varied permeate flux were less significant for the other three N-nitrosamines (Fig. 268 6). The results here indicate that the heat-treated Prototype membrane is resilient to changes 269 in operating conditions, which is important to achieve a stable operation. Although permeate 270 flux is usually unchanged during RO system operation, there could be a large variation in 271 permeate flux inside a vessel of a three-stage RO where permeate flux at later stages is 272 typically very low (Hoek et al., 2008). A multiple-stage configuration is commonly utilized in 273 water recycling to achieve a recovery near 85% (Fujioka et al., 2012a). Thus, if the 274 membrane were to be implemented at a full scale, the high stable rejection under variable
275 permeate flux would be advantageous in maintaining high NDMA rejection throughout the
276 RO train.

277

#### [Fig. 6]

## 278 *3.3.* Fouling propensity

279 The fouling propensity of the ESPA2 and heat-treated Prototype membranes was evaluated 280 with a UF-treated wastewater effluent. Initial TMPs for the ESPA2 and Prototype membrane 281 were 0.77 and 3.9 MPa, respectively. Operation of the membranes at a high permeate flux 282 resulted in progressive fouling of the ESPA2 membrane. Permeability dropped to 87% of the 283 initial level after 23 h (Fig. 7). In contrast, the permeability of the heat-treated Prototype 284 membrane remained at 98% of the initial level between 0.5-23 h. The mechanism of the 285 improved fouling resistance by heat treatment has not been determined and will be the scope 286 of future studies.

287

#### [Fig. 7]

## 288 *3.4. Outlook for high rejection RO membranes*

289 One of the drawbacks of high rejection RO membranes is low permeability. Lower 290 membrane permeability equates to higher TMP to produce comparable permeate flux at the 291 expense of high energy cost. Nevertheless, greater NDMA removal capability by RO could 292 potentially lead to less energy utilization by the UV-based AOP. UV/AOP requires 293 ~1,000 mJ/cm<sup>2</sup> to provide a 1.2-log reduction for NDMA (Sharpless et al., 2003) and the 294 energy requirement is about 10 times greater than a typical UV dose for a 4-log removal of 295 Cryptosporidium (USEPA, 2006). Energy consumption of AOP could be cut in half if 0.6-log 296 removal credits for NDMA were given to the RO process with only 0.6-log removal needed for NDMA by the UV/AOP. Nevertheless, because the increase in feed pressure requires a 297

greater capital investment in RO feed pumps and greater energy consumption, the impact of a reduction in energy consumption by AOP on the overall energy requirement still remains unknown. Thus, a feasibility investigation is vital to the development of a new RO system configuration with high rejection RO membranes and high pressure pumps. More importantly, the development of a high rejection RO membrane with high permeability is needed to improve the security of recycled water in potable reuse and make potable water reuse more cost-effective.

## 305 4. Conclusions

306 Heat treatment was applied to three commercial and one prototype RO membrane to develop 307 a high rejection membrane with over 1.0-log NDMA removal. Among four untreated and 308 four heat-treated RO membranes, the heat-treated Prototype membrane was able to achieve a 309 1.1-log (92%) removal of NDMA. Nevertheless, permeability was considerably reduced by 310 heat treatment. DMA, an NDMA precursor, was well rejected by the heat-treated Prototype 311 membrane (>98%), and near complete rejection of high molecular weight secondary amines 312 (i.e., MOR, PYR and PIP) was achieved. In addition, rejection by the heat-treated Prototype 313 membrane remained stable under conditions of changing feed temperature and permeate flux. 314 When operated on UF-treated wastewater, permeability of the membrane remained high (98%), as compared to the ESPA2 membrane (87%). Despite the very low permeability of the 315 316 heat-treated Prototype membrane, the results indicate that its enhanced ability to reject N-317 nitrosamines could contribute to improved security of recycled water in potable water reuse. 318 Further development of highly selective RO membranes with high permeability is still needed 319 to ensure the feasibility of using these membranes at full scale.

## 320 5. Acknowledgements

321 We thank Hydranautics/Nitto for providing RO membrane samples for this investigation.

### 322 **6. References**

- Arnold, R.G., Sáez, A.E., Snyder, S., Maeng, S.K., Lee, C., Woods, G.J., Li, X. and Choi, H.
  (2012) Direct potable reuse of reclaimed wastewater: it is time for a rational discussion. Rev. Environ. Health 27(4), 197-206.
- Bellona, C., Budgell, K., Ball, D., Spangler, K., Drewes, J.E. and Chellam, S. (2011) Models
  to predict organic contaminant removal by RO and NF membranes. IDA J. 3(2), 4044.
- Bellona, C., Drewes, J.E., Oelker, G., Luna, J., Filteau, G. and Amy, G. (2008) Comparing
   nanofiltration and reverse osmosis for drinking water augmentation. J. AWWA 100(9),
   102-116.
- Bellona, C., Drewes, J.E., Xu, P. and Amy, G. (2004) Factors affecting the rejection of
  organic solutes during NF/RO treatment A literature review. Water Res. 38(12),
  2795-2809.
- Bernstein, R., Belfer, S. and Freger, V. (2011) Toward Improved Boron Removal in RO by
   Membrane Modification: Feasibility and Challenges. Environ. Sci. Technol. 45(8),
   3613-3620.
- Burgess, J., Meeker, M., Minton, J. and O'Donohue, M. (2015) International research agency
   perspectives on potable water reuse. Environ. Sci.: Water Res. Technol. 1(5), 563-580.
- CSWRCB (2016) Investigation on the feasibility of developing uniform water recycling
   criteria for direct potable reuse. California State Water Resources Control Board.
- 342 Debroux, J.-F., Soller, J.A., Plumlee, M.H. and Kennedy, L.J. (2012) Human Health Risk
  343 Assessment of Non-Regulated Xenobiotics in Recycled Water: A Review. Hum. Ecol.
  344 Risk Assess.: Int. J. 18(3), 517-546.
- Dražević, E., Košutić, K. and Freger, V. (2014) Permeability and selectivity of reverse
  osmosis membranes: Correlation to swelling revisited. Water Res. 49(0), 444-452.
- Drewes, J.E. and Khan, S.J. (2015) Contemporary design, operation, and monitoring of
   potable reuse systems. J. Water Reuse Desalination 5(1), 1-7.
- Farré, M.J., Döderer, K., Hearn, L., Poussade, Y., Keller, J. and Gernjak, W. (2011a)
  Understanding the operational parameters affecting NDMA formation at Advanced
  Water Treatment Plants. J. Hazard. Mater. 185(2-3), 1575-1581.
- Farré, M.J., Keller, J., Holling, N., Poussade, Y. and Gernjak, W. (2011b) Occurrence of N nitrosodimethylamine precursors in wastewater treatment plant effluent and their fate

- during ultrafiltration-reverse osmosis membrane treatment. Water Sci. Technol. 63(4),
   605-612.
- Fleming, E.C., Pennington, J.C., Wachob, B.G., Howe, R.A. and Hill, D.O. (1996) Removal
   of N-nitrosodimethylamine from waters using physical-chemical techniques. J.
   Hazard. Mater. 51(1-3), 151-164.
- 359 Freger, V. (2015) Outperforming nature's membranes. Science 348(6241), 1317-1318.
- Fujioka, T., Khan, S.J., McDonald, J.A., Henderson, R.K., Poussade, Y., Drewes, J.E. and
   Nghiem, L.D. (2013a) Effects of membrane fouling on *N*-nitrosamine rejection by
   nanofiltration and reverse osmosis membranes. J. Membr. Sci. 427, 311-319.
- Fujioka, T., Khan, S.J., McDonald, J.A., Roux, A., Poussade, Y., Drewes, J.E. and Nghiem,
  L.D. (2013b) *N*-nitrosamine rejection by reverse osmosis membranes: A full-scale
  study. Water Res. 47(16), 6141-6148.
- Fujioka, T., Khan, S.J., Poussade, Y., Drewes, J.E. and Nghiem, L.D. (2012a) *N*-nitrosamine
  removal by reverse osmosis for indirect potable water reuse A critical review based
  on observations from laboratory-, pilot- and full-scale studies. Sep. Purif. Technol. 98,
  503-515.
- Fujioka, T., Nghiem, L.D., Khan, S.J., McDonald, J.A., Poussade, Y. and Drewes, J.E.
  (2012b) Effects of feed solution characteristics on the rejection of *N*-nitrosamines by
  reverse osmosis membranes. J. Membr. Sci. 409–410, 66-74.
- Fujioka, T., Oshima, N., Suzuki, R., Higgins, M., Price, W.E., Henderson, R.K. and Nghiem,
  L.D. (2015) Effect of heat treatment on fouling resistance and the rejection of small
  and neutral solutes by reverse osmosis membranes. Water Sci. Technol. Water Supply
  15(3), 510-516.
- Fujioka, T., Takeuchi, H., Tanaka, H., Nghiem, L.D., Ishida, K.P. and Kodamatani, H. (2016)
  A rapid and reliable technique for *N*-nitrosodimethylamine analysis in reclaimed
  water by HPLC-photochemical reaction-chemiluminescence. Chemosphere 161, 104111.
- Geise, G.M., Park, H.B., Sagle, A.C., Freeman, B.D. and McGrath, J.E. (2011) Water
   permeability and water/salt selectivity tradeoff in polymers for desalination. J. Membr.
   Sci. 369(1-2), 130-138.
- Hoek, E.M.V., Allred, J., Knoell, T. and Jeong, B.-H. (2008) Modeling the effects of fouling
  on full-scale reverse osmosis processes. J. Membr. Sci. 314(1-2), 33-49.
- Hofs, B., Schurer, R., Harmsen, D.J.H., Ceccarelli, C., Beerendonk, E.F. and Cornelissen,
  E.R. (2013) Characterization and performance of a commercial thin film
  nanocomposite seawater reverse osmosis membrane and comparison with a thin film
  composite. J. Membr. Sci. 446, 68-78.
- Kodamatani, H., Iwaya, Y., Saga, M., Saito, K., Fujioka, T., Yamazaki, S., Kanzaki, R. and
  Tomiyasu, T. (2017) Ultra-sensitive HPLC-photochemical reaction-luminol
  chemiluminescence method for the measurement of secondary amines after
  nitrosation. Anal. Chim. Acta 952, 50-58.

- Kodamatani, H., Yamasaki, H., Sakaguchi, T., Itoh, S., Iwaya, Y., Saga, M., Saito, K.,
  Kanzaki, R. and Tomiyasu, T. (2016) Rapid method for monitoring Nnitrosodimethylamine in drinking water at the ng/L level without pre-concentration
  using high-performance liquid chromatography-chemiluminescence detection. J.
  Chromatogr. A 1460, 202-206.
- Krauss, M., Longrée, P., van Houtte, E., Cauwenberghs, J. and Hollender, J. (2010)
  Assessing the fate of Nitrosamine precursors in wastewater treatment by
  physicochemical fractionation. Environ. Sci. Technol. 44(20), 7871-7877.
- 402 Leverenz, H.L., Tchobanoglous, G. and Asano, T. (2011) Direct potable reuse: A future
  403 imperative. J. Water Reuse Desalination 1(1), 2-10.
- 404 McCurry, D.L., Ishida, K., Oelker, G. and Mitch, W.A. (2017) Reverse Osmosis Shifts
  405 Chloramine Speciation Causing Re-Formation of NDMA during Potable Reuse of
  406 Wastewater. Environ. Sci. Technol. 51(15), 8589-8596.
- Mitch, W.A., Gerecke, A.C. and Sedlak, D.L. (2003) A N-Nitrosodimethylamine (NDMA)
  precursor analysis for chlorination of water and wastewater. Water Res. 37(15), 37333741.
- Miyashita, Y., Park, S.-H., Hyung, H., Huang, C.-H. and Kim, J.-H. (2009) Removal of NNitrosamines and their precursors by nanofiltration and reverse osmosis membranes. J.
  Environ. Eng. 135(9), 788-795.
- Plumlee, M.H., López-Mesas, M., Heidlberger, A., Ishida, K.P. and Reinhard, M. (2008) Nnitrosodimethylamine (NDMA) removal by reverse osmosis and UV treatment and
  analysis via LC-MS/MS. Water Res. 42(1-2), 347-355.
- Poussade, Y., Roux, A., Walker, T. and Zavlanos, V. (2009) Advanced oxidation for indirect
  potable reuse: a practical application in Australia. Water Sci. Technol. 60(9), 24192424.
- Schmidt, C.K. and Brauch, H.-J. (2008) N,N-Dimethylsulfamide as precursor for NNitrosodimethylamine (NDMA) formation upon ozonation and its fate during
  drinking water treatment. Environ. Sci. Technol. 42(17), 6340-6346.
- Schreiber, I.M. and Mitch, W.A. (2005) Influence of the order of reagent addition on NDMA
  formation during chloramination. Environ. Sci. Technol. 39(10), 3811-3818.
- 424 Sgroi, M., Roccaro, P., Oelker, G.L. and Snyder, S.A. (2015) N-nitrosodimethylamine 425 (NDMA) formation at an indirect potable reuse facility. Water Res. 70, 174-183.
- Shah, A.D. and Mitch, W.A. (2011) Halonitroalkanes, halonitriles, haloamides, and *N*Nitrosamines: A critical review of nitrogenous disinfection byproduct formation pathways. Environ. Sci. Technol. 46(1), 119-131.
- Sharpless, C.M., Page, M.A. and Linden, K.G. (2003) Impact of hydrogen peroxide on nitrite
   formation during UV disinfection. Water Res. 37(19), 4730-4736.

- Shintani, T., Matsuyama, H. and Kurata, N. (2009) Effect of heat treatment on performance
  of chlorine-resistant polyamide reverse osmosis membranes. Desalination 247(1–3),
  370-377.
- Soltermann, F., Lee, M., Canonica, S. and von Gunten, U. (2013) Enhanced *N*-nitrosamine
  formation in pool water by UV irradiation of chlorinated secondary amines in the
  presence of monochloramine. Water Res. 47(1), 79-90.
- 437 Steinle-Darling, E., Zedda, M., Plumlee, M.H., Ridgway, H.F. and Reinhard, M. (2007)
  438 Evaluating the impacts of membrane type, coating, fouling, chemical properties and
  439 water chemistry on reverse osmosis rejection of seven nitrosoalklyamines, including
  440 NDMA. Water Res. 41(17), 3959-3967.
- 441 USEPA (1993) N-Nitrosodimethylamine (CASRN 62-75-9) Integrated risk information
   442 system (IRIS), U.S. Environmental Protection Agency.
- 443 USEPA (2006) Ultraviolet disinfection guidance manual for the final long term 2 enhanced
   444 surface water treatment rule. Office of Water (4601) EPA 815-R-806-007.
- Van der Bruggen, B., Verliefde, A., Braeken, L., Cornelissen, E.R., Moons, K., Verberk,
  J.Q.J.C., van Dijk, H.J.C. and Amy, G. (2006) Assessment of a semi-quantitative
  method for estimation of the rejection of organic compounds in aqueous solution in
  nanofiltration. J. Chem. Technol. Biotechnol. 81(7), 1166-1176.
- Werber, J.R., Deshmukh, A. and Elimelech, M. (2016a) The Critical Need for Increased
   Selectivity, Not Increased Water Permeability, for Desalination Membranes. Environ.
   Sci. Technol. Letters 3(4), 112-120.
- Werber, J.R., Osuji, C.O. and Elimelech, M. (2016b) Materials for next-generation
   desalination and water purification membranes. Nature Reviews: Materials 1, 16018.
- Zhang, H. and Geise, G.M. (2016) Modeling the water permeability and water/salt selectivity
   tradeoff in polymer membranes. J. Membr. Sci. 520, 790-800.

Compound	Molecular	Molecular	Log D	$pK_a{}^a$	Minimum
	formula	weight [Da]	at pH 8 <sup>a</sup>		projection area <sup>a</sup>
					[Å]
N-Nitrosamines					
NDMA	$C_2H_6N_2O$	74.1	0.04	3.5	19.5
NMEA	$C_2H_8N_2O$	88.1	0.40	3.4	21.9
NPYR	$C_4H_8N_2O$	100.1	0.44	3.3	25.0
NDEA	$C_4H_{10}N_2O$	102.1	0.75	3.3	25.4
NPIP	$C_5H_{10}N_2O$	114.1	0.89	3.3	27.2
NMOR	$C_4H_8N_2O_2$	116.1	-0.18	3.1	25.2
Secondary amines					
DMA	C <sub>2</sub> H <sub>7</sub> N	45.1	-2.64	10.5	15.8
PYR	C4H9N	71.1	-2.80	11.4	22.5
DEA	$C_4H_{11}N$	73.1	-1.98	10.6	20.3
PIP	$C_5H_{11}N$	85.2	-1.69	10.4	25.6
MOR	C4H9NO	87.1	-1.03	8.5	21.5

**Table 1** – Physicochemical properties of the selected N-nitrosamines and secondary amines.

<sup>a</sup> Chemicalize (<u>http://www.chemicalize.org</u>).

Membrane	Treatment	Pure water	Conductivity	Zeta potential
		permeability*	rejection*	at pH 7.9**
		[L/m <sup>2</sup> hbar]	[%]	[mV]
ESPA2	untreated	$6.3\pm0.2$	$99.0\pm0.0$	-51 ± 11
	heat-treated	$5.0\pm0.2$	$98.8\pm0.3$	$-44 \pm 8$
ESPAB	untreated	$4.3\pm0.1$	$98.5\pm0.1$	$-28 \pm 3$
	heat-treated	$2.9\pm0.04$	$97.3\pm0.1$	$-35 \pm 2$
HYDRA	untreated	$1.9\pm0.03$	$99.3\pm0.03$	$-18 \pm 1$
	heat-treated	$1.2\pm0.03$	$98.9\pm0.2$	-21 ± 2
Prototype	untreated	$1.5\pm0.02$	$99.8\pm0.02$	$-33 \pm 4$
	heat-treated	$1.0\pm0.04$	$99.5\pm0.1$	-37 ± 2

**Table 2** – Properties of the selected RO membranes.

\*Pure water permeability and conductivity rejection were obtained from two replicates.

\*\*Zeta potential was obtained from three replicates.

## **FIGURES**

Fig. 1 – Rejection of NDMA and conductivity by untreated and heat-treated RO membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, feed temperature =  $20.0 \pm 0.1$  °C, permeate flux =  $20 \text{ L/m}^2\text{h}$ ). Error bars show the range of two replicate experiments.

**Fig. 2** – Correlation between NDMA rejection and pure water permeability of eight RO membranes. Error bars show the range of two replicate experiments.

**Fig. 3** – Correlation between minimum projected area and the rejection of *N*-nitrosamine by (a) untreated ESPA2 and (b) heat-treated Prototype membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, permeate flux = 20 L/m<sup>2</sup>h, feed temperature =  $20.0 \pm 0.1$  °C). Error bars show the range of two replicate experiments.

**Fig. 4** – Correlation between minimum projected area and the rejection of secondary amines by (a) untreated ESPA2 and (b) heat-treated Prototype membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, permeate flux = 20 L/m<sup>2</sup>h, feed temperature =  $20.0 \pm 0.1$  °C). Solid symbol indicates that the permeate concentration was below the instrumental detection limit. Error bars show the range of two replicate experiments.

**Fig. 5** – Effects of feed temperature on *N*-nitrosamine rejection by (a) untreated ESPA2 and (b) heat-treated Prototype membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, permeate flux =  $20 \text{ L/m}^2\text{h}$ ).

**Fig. 6** – Effects of permeate flux on *N*-nitrosamine rejection by (a) untreated ESPA2 and (b) heat-treated Prototype membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, feed temperature =  $20.0 \pm 0.1$  °C).

**Fig. 7** – RO treatment using a UF-filtered wastewater by (a) untreated ESPA2 and (b) heattreated Prototype membranes (permeate flux = 40 L/m<sup>2</sup>h, feed temperature =  $20.0 \pm 0.1$  °C). The initial TMP for EPSA2 and heat-treated Prototype membranes were 0.77 and 3.9 MPa, respectively.



Fig. 1



Fig. 2











Fig. 5



Fig. 6



Fig. 7

# High rejection reverse osmosis membrane for removal of *N*-nitrosamines and their precursors

3 Takahiro Fujioka<sup>1,\*</sup>, Kenneth P. Ishida<sup>2</sup>, Takuji Shintani<sup>3</sup>, Hitoshi Kodamatani<sup>4</sup>

4	<sup>1</sup> Water and Environmental Engineering, Graduate School of Engineering,
5	Nagasaki University, 1-14 Bunkyo-machi, Nagasaki 852-8521, Japan
6	<sup>2</sup> Research & Development Department, Orange County Water District, 18700 Ward Street,
7	Fountain Valley, CA 92708, USA
8	<sup>3</sup> Division of Advanced Membrane Science and Technology, Graduate School of Science,
9	Technology and Innovation, Kobe University, 1-1 Rokkodai-cho, Kobe 657-8501, Japan
10	<sup>4</sup> Division of Earth and Environmental Science, Graduate School of Science and Engineering,
11	Kagoshima University, 1-21-35 Korimoto, Kagoshima 890-0065, Japan

# SUPPLEMENTARY MATERIAL

13

<sup>14 \*</sup> Corresponding author: Takahiro Fujioka, Email: tfujioka@nagasaki-u.ac.jp, Ph +81 095 819 2695

N-nitrosamines	Structure <sup>a</sup>	Secondary amine	Structure <sup>a</sup>
NDMA		DMA	H₃C <sup>∕NH</sup> CH₃
NMEA			
NPYR		PYR	HN
NDEA		DEA	H <sub>3</sub> C
NPIP		PIP	HN
NMOR		MOR	

 Table S1 – Structure of the selected N-nitrosamines and secondary amines.

<sup>a</sup>Chemspider (http://www.chemspider.com/).



**Fig. S2** – Schematic figure of minimum projection area. The line perpendicular to the circular disk represents the centre axis of the minimum projection area.



Fig. S3 – Schematic diagram of the cross-flow RO filtration system.



Fig. S4 – TMP and rejection of NMEA, NPYR and NMOR by untreated and heat-treated RO membranes (20 mM NaCl, 1 mM NaHCO<sub>3</sub>, 1 mM CaCl<sub>2</sub>, feed temperature =  $20.0 \pm 0.1$  °C, permeate flux =  $20 \text{ L/m}^2\text{h}$ ).