1	Hydrogeochemical evolution of groundwater in a Quaternary sediment and Cretaceous
2	sandstone unconfined aquifer in Northwestern China
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17	Abstract
18	A better understanding of the hydrogeochemical evolution of groundwater in vulnerable aquifers is
19	important for the protection of water resources. To assess groundwater chemistry, groundwater
20	sampling was performed from different representative aquifers in 2012-13. A Piper trilinear diagram
21	showed that the groundwater types can be classified into Na-SO4 and Na-Cl types. Only one
22	groundwater sample was Na-HCO3 type. The dominant cations for all samples were Na ⁺ . However,

23	the dominant anions varied from HCO_3^- to SO_4^{2-} , and as well Cl ⁻ . The mean total dissolved solid
24	(TDS) content of groundwater in the region was 1,889 mg/L. Thus, only 20% of groundwater
25	samples meet Chinese drinking water standards (<1,000 mg/L). Principal component analysis
26	(PCA) combined with hierarchical cluster analysis (HCA) and self-organizing maps (SOM) were
27	applied for the classification of the groundwater geochemistry. The three first principal components
28	explained 58, 20, and 16% of the variance, respectively. The first component reflects sulfate
29	minerals (gypsum, anhydrite) and halite dissolution, and/or evaporation in the shallow aquifer. The
30	second and third components are interpreted as carbonate rock dissolution. The reason for two
31	factors is that the different aquifers give rise to different degree of hydrogeochemical evolution
32	(different travel distances and travel times). Identified clusters for evolution characteristic and
33	influencing factors were confirmed by the PCA-HCA methods. Using information from eight ion
34	components and SOM, formation mechanisms and influencing factors for the present groundwater
35	quality were determined.
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37	Keywords
38	Self-organizing maps, Hydrogeochemical characteristics, The Dosit River, Principal component
39	analysis, Sulfate minerals
40	
41	Introduction
42	The Ordos Basin is part of a large-scale sedimentary geological area in northwestern China.
43	Abundant organic minerals such as coal, petroleum, natural gas, and halite have been deposited in
44	the basin, which makes it one of the largest sources for energy and petrochemical production in

45	China (Jiang et al., 2012). During recent decades, the economy and mining industry have boomed
46	in the area, resulting in a dramatic increase in water demand. However, due to the lack of surface
47	water, groundwater is the main water resource for local development needs. The Dosit River is one
48	of a number of the sub-basins of the Ordos Basin with similar resources and water problems. In
49	order to meet the water resources demands for drinking water and industrial water supply in this
50	region, research has focused on determining: (1) the groundwater hydrogeochemical characteristics;
51	(2) recharge and evaporation amount estimates; and (3) assessment of the characteristics of the local
52	groundwater flow systems (Yin et al., 2011; Jiang et al., 2014; Wang et al., 2015, Jiang et al., 2018).
53	As water scarcity is a serious issue in the region, determining the key aspects of the hydrogeology
54	of the region is important for the local economic development. This is particularly the case for
55	understanding the quality of groundwater and its availability for extractive uses, the characteristics
56	of surface-groundwater interactions, and for an improved understanding of the hydrogeochemistry
57	and groundwater circulation characteristics. To improve this understanding, groundwater samples
58	were collected in the downstream region of the Dosit River for assessing water quality and its
59	chemical composition. Piper trilinear diagrams, multivariate statistical analyses such as principal
60	component analysis (PCA) and hierarchical cluster analysis (HCA) were used to investigate
61	collected groundwater samples (e.g., Brown, 1998; Cloutier et al., 2008; Nakagawa et al., 2016).
62	Recently, self-organizing maps (SOM) have been shown to efficiently classify groundwater
63	chemistry (Choi et al., 2014; Nguyen et al., 2015). Consequently, we combined SOM with PCA-
64	HCA methods to investigate groundwater chemistry for the downstream Dosit River area.

In view of the above, the main objectives of this study were: (1) to assess groundwater quality, hydrogeochemistry, and evolution characteristics of the groundwater by use of the aforementioned techniques; and (2) to determine usefulness of these methods for hydrogeochemical
investigations in regions like the Dosit River area.

70 Materials and methods

71 Study area

The Dosit River Basin is located between 106°54'28" and 108°16'16" E, and 38°18'21" and 39° 36'06'' N (Fig. 1). It occupies an area of 10,924 km² and constitutes a sub-basin of the Ordos Basin in Northwestern China. In turn, the Dosit River discharges into the important Yellow River. The altitude of the Dosit River Basin varies from 1,080 to 1,500 m. The basin is mostly surrounded by mountains with the exception of the west and water flows in a westerly direction in the basin. The average monthly temperatures range from -10.5 °C in January to 22.4 °C in July and the mean annual temperature is about 6.9 °C (1955-2007). The mean annual rainfall of the region is about 267 mm (1955-2007) and 60-80% of rainfall takes place from June to September. The mean annual evaporation is about 2,465 mm (Sun, 2010; Jiang et al., 2014). The basin climate is thus characterized as arid to semiarid.

The aquifers of the Dosit River watershed can be classified into two groups. The first, is the uppermost aeolian and alluvial-lacustrine pore aquifer system in Quaternary sediments that is a thin and uneven unconfined aquifer. The second, deeper aquifer system is constituted by a poorly consolidated, pore-fissure aquifer system of Cretaceous sandstone with a thickness of 700-1,000 m which is the main aquifer system of the Dosit River Basin. The Jurassic mudstone with coal layers, which underlies the Cretaceous sandstone, is generally assumed to be an aquiclude. The main minerals of the Cretaceous sandstone aquifer contain quartz, albite, and feldspar, as well as some minor minerals such as gypsum, mirabilite, halite, calcite, and dolomite (Sun, 2010; Wang et al., 2015). According to hydrogeochemical and isotope studies, a three-part structure with boundaries at depths of 200 m and 600-750 m, is consistent with local, intermediate, and regional groundwater flow systems (Wang et al., 2015). The principle recharge and discharge areas of local flow systems are mainly distributed on sides of valley slopes. Groundwater discharge from the deeper aquifers mostly takes place in river channels and recharge to these aquifers mostly takes place at a distance from major rivers. The groundwater of these flow systems contributes to baseflow in the Dosit River (Wang et al., 2016).

Sampling and analysis

Groundwater samples were collected by pumping at 20 locations from existing water supply wells in the downstream part of the Dosit River in 2012-2013 (Fig. 1). Water samples were collected in pre-washed bottles. Temperature, pH, and electrical conductivity (EC) were analyzed at the sampling sites, CO_3^{2-} and HCO_3^{-} concentrations were identified with titration by use of HCl. Cl⁻ and SO₄²⁻ concentration were analyzed by ion chromatography (ICP-900, Dionex), while Na⁺, K^+ , Mg^{2+} , and Ca^{2+} concentrations were measured by inductively coupled plasma (ICP-900, Thermo) (Wang et al., 2015). Charge balance errors (CBE) were checked for all groundwater samples according to:

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$$CBE = \frac{\sum \text{cations} - \sum \text{anions}}{\sum \text{cations} + \sum \text{anions}} \times 100$$

where all ions concentrations are expressed in mmol_c/L. We confirmed that all CBEs of samples were less than 10%.

111 Multivariate statistical analysis

Principal component analysis (PCA) is a well-tested multivariate statistical method that has been widely used to analyze hydrochemical groundwater data. Its key feature is data reduction from high- to low-dimensional space (Morell et al., 1996; Choi et al., 2014). It extracts synthetic variables with minimal information loss (Aiuppa et al., 2003; Nakagawa et al., 2017). HCA is an effective method for data classification. It has been used to identify clusters of groundwater samples based on similarity of hydrogeochemical components (Cloutier et al., 2008; Montcoudiol et al., 2015). The classification result is expressed in a dendrogram, using the Euclidean distance as a distance measure (Ward's method) (Güler et al., 2002). In this study, the statistical software JMP Pro 13 (SAS Institute Inc.) was used when using PCA and HCA.

The Self-organizing map (SOM) technique is also a powerful and effective tool for data classification (Nguyen et al., 2015; Nakagawa et al., 2017). It has been used in different research fields such as hydrology (Kalteh and Berndtsson, 2007), wastewater treatment (Yu et al., 2014), and meteorology (Nishiyama et al., 2007). As well, SOM has been applied to investigating water chemistry of river and groundwater (Choi et al., 2014; Nguyen et al., 2015). SOM is a type of artificial neural network technique, which is distinguished by unsupervised training (Kohonen, 2001). It can project high-dimensional data onto a low-dimensional array, and let complex target data simplify into a regular arranged map based on the degree of similarity (Jin et al., 2011). In general, the purpose of the SOM application is to acquire useful and informative reference vectors. These vectors are obtained by iterative updates in the training phase of SOM that is made up from three main steps; competition between nodes, selection of a winner nodes, and updating of the reference vectors (Nguyen et al., 2015). When applying SOM methodology, the selection of an

appropriate initialization and data transformation method is important. On the basis of SOM properties, larger map sizes will obtain a higher resolution for pattern recognition. The optimal number of map nodes is determined by heuristic rules according to $m = 5\sqrt{n}$, where *m* represents the number of map nodes and n represents the number of input data (Hentati et al., 2010). The number of rows and columns is dependent on square root of the ratio between the two largest eigenvalues of the transformed data (García and González, 2004). The eigenvalues are calculated by PCA.

Following the above rules, the SOM structure was organized. Using a linear initialization technique, each node is set with a reference vector. Under the proper limited data condition, a linear initialization technique is better for the pattern classification as compared to random initialization due to small data sets and boundary effects (Nguyen et al., 2015). More details of SOM are discussed by Kohonen, (1982, 2001) and Vesanto et al. (2000). Results of SOM analysis are achieved at the end of the training process, which is fine-tuned using cluster analysis. K-means algorithms, which is a partitioned algorithm, is frequently used in SOM (Jin et al., 2011). Davies-Bouldin Index (DBI) applying k-means algorithms determine the optimal number of clusters (García and González, 2004). According to the principle of "similarity within a cluster" and "dissimilarity between clusters", the DBI values were calculated from the minimum to maximum number of clusters. The smaller DBI values show that the dissimilarity of each cluster becomes larger (Nakagawa et al., 2017). In other words, the minimum DBI corresponds to the optimal number of clusters in SOM application. These processes were put into practice using a modified version of SOM Toolbox 2.0 (Vesanto et al., 2000).

Results and discussion

156 Hydrogeochemical characteristics

Table 1 shows the statistical summary of hydrogeochemical variables (Na⁺, K⁺, Ca²⁺, Mg²⁺, Cl⁻, SO₄²⁻, CO₃²⁻, HCO₃⁻, TDS, pH) for the 20 samples. Mean TDS corresponded to 1,889 mg/L, with a minimum of 455 mg/L and a maximum of 3,974 mg/L. TDS concentration of surface water is generally higher in the downstream of Dosit River, due to evaporation and the effects of groundwater evolution. pH varied from 8.0 to 9.0, indicating that the water environment is slightly alkaline. Figure 2 shows results plotted in a Piper trilinear diagram (Piper, 1944). All samples are located in area IV. Only one sample is classified as Na-HCO₃ type. Seven samples showed Na-SO₄ type. The other samples represented Na-Cl type. Cations were dominated by Na⁺ in all samples. Principal anions changed from HCO_3^- to SO_4^{2-} , and then to Cl^- along groundwater flow direction. This process corresponds to anion evolution according to $HCO_3^- \rightarrow HCO_3^- + SO_4^{2-} \rightarrow SO_4^{2-} + Cl^- \rightarrow Cl^ +SO_4^2 \rightarrow Cl^-$ from recharge to discharge areas in aquifers of large sedimentary basins (Chebotarev, 1955; Singhal and Gupta 2010).

170 Principal component and hierarchical cluster analysis

Eight parameters (Na⁺, K⁺, Mg²⁺, Ca²⁺, Cl⁻, HCO₃⁻, CO₃²⁻, and SO₄²⁻) were selected for analysis by PCA and HCA. Firstly, input data were standardized, and then eigenvalues, factor loadings, and principal component scores were calculated using a correlation matrix (Nakagawa et al., 2016). The total number of common factors in the PCA was selected on the basis of the Kaiser Criterion (Cloutier et al., 2008). In this criterion, if an eigenvalue is greater than 1, it is retained. According to this rule, the first three components (Factors 1, 2, and 3) were extracted. Ward's

177	method with Euclidean distances was selected for HCA. For more distinct understanding of the
178	eight ionic correlations, TDS, pH, and water table depth were not considered in the PCA-HCA
179	analysis. The result of PCA is shown in Table 2. The first three principal components accounted for
180	a total of 94% of the variance. The three components explained 57.8, 20.1, and 16.1% of the
181	variance, respectively. Hereafter, the first three principal components are called Factors 1, 2 and 3,
182	respectively. Factor 1 had positive loadings for Ca ²⁺ , Mg ²⁺ , SO ₄ ²⁻ , Na ⁺ , Cl ⁻ , and K ⁺ ($r = 0.80 \sim 0.94$).
183	Factor 1 is interpreted as sulfate minerals (gypsum and anhydrite) and halite dissolution, and/or
184	evaporation in the shallow aquifer. Factor 2 has positive high loadings for HCO_3^- , and CO_3^{2-} ($r =$
185	0.66 and 0.93). Factor 2 is mainly affected by carbonate rock dissolution. Factor 3 has moderately
186	positive loadings for Mg ²⁺ and Ca ²⁺ ($r = 0.49$ and 0.50). Factor 3 is also mainly affected by carbonate
187	rock dissolution. The reason may be a different degree of carbonate rock dissolution due to different
188	travel distances and travel times in the different aquifers.

The scatter plot of the 20 samples described by principal components (Factors 1 and 2; Factors 2 and 3) and classified into four clusters based on HCA is shown in Fig. 3. If a factor score is greater than 0, it means that the component is influenced by the water chemistry characteristic at the site. Conversely, if a factor score is less than 0, it means that the component was not significantly affected by the water chemistry at the site (Banoeng-Yakubo et al., 2009). In Fig. 3(a) Cluster A is to some extent influenced by Factor 2. Cluster C is to some extent affected by Factor 1. Clusters B and D are influenced by both Factors 1 and 2. As well, Fig. 3(b) indicates that Clusters A and B are affected by Factor 3. Factor 3 has less influence on Clusters C and D. According to the local and regional groundwater flow systems of the Dosit River Watershed (Wang et al., 2016) and evolution, samples of Clusters A and B are located at different sites but in the same aquifer (depth less than

200 m; local flow system). Samples of Cluster A are located in the upstream area as opposed to that
of Cluster B (Fig. 4). Similarly, samples of Clusters C and D are located at different sites but in the
same aquifer (depth between 200 and 600 m; regional flow system). The location of samples
classified into Cluster D is closer to the river than that of Cluster C (Fig. 4).

Water samples were classified into Clusters A, B, C, and D in a Piper trilinear diagram (Fig. 2). Water composition of samples in Cluster A was modified to that of Cluster B due to sulfate minerals and halite dissolution with groundwater flow from upstream to downstream (Figs. 2 and 4). Because of the groundwater discharge from the peripheral aquifer to the river (Jiang et al., 2018), carbonate dissolution changed water composition of samples in Cluster C to that in Cluster D (Figs. 2 and 4). Thus, groundwater samples of Cluster A affected by Factor 1 appear to evolve into groundwater samples of Cluster B. As well, water samples of Cluster C affected by Factor 2 appear to evolve into water samples of Cluster D.

212 Self-organizing map analysis

Based on the methodology described above, concentrations of the eight chemical variables $(Na^+, K^+, Ca^{2+}, Mg^{2+}, Cl^-, SO_4^{2-}, CO_3^{2-}, and HCO_3^{-})$ for the 20 samples were used as input to the SOM application. Figure 5 shows visual SOMs for each parameter after the training process. The SOMs are characterized by 24 nodes (number of rows and columns are 6 and 4, respectively) and display concentration characteristics of each variable. If the node color is dark gray, it represents variables that have high values. On the contrary, nodes color that are light gray, means that variables have low values. From Fig. 5 it can be seen that maps for Ca^{2+} , Mg^{2+} , and SO_4^{2-} have similar gray gradients, indicating that there is a strong positive correlation among these ions. Maps for Na⁺ and

221	Cl ⁻ also have a similar gray gradient. In contrast, HCO ₃ ⁻ displays negative correlation with inverse
222	gray gradients for the SOM maps. The main sources of HCO3 ⁻ is dissolution of carbonate rock
223	(calcite, dolomite), which commonly occurs in the area (Sun, 2010). However, there is no significant
224	correlation between HCO_3^- and cations, which suggests that ion exchange of cations and carbonate
225	precipitation occurs. Chloro-alkaline indices (CAI-I and CAI-II) suggested by Schoeller (1972)
226	were calculated to assess the ion exchange. Both CAI-I and CAI-II for all samples were negative
227	from -2.99 to -0.15, and -1.02 to -0.13, respectively, indicating that there is exchange Na^+ and K^+ in
228	the soil matrix with Ca^{2+} and Mg^{2+} in groundwater. Saturation Indices (SI) with respect to calcite,
229	dolomite, and aragonite were calculated by using PhreeqcI version 3.20-9820 in this study
230	(Parkhurst and Appelo 2013). The SI values of calcite, dolomite, and aragonite ranged from -0.33
231	to 1.38, -1.49 to 2.93, and -0.47 to 1.24, respectively. Except for the sampling location CBS3 that
232	is located in the upstream reaches, most of the samples showed positive SI values, which explain
233	that groundwater is saturated to oversaturated with these carbonate minerals in the lower reaches of
234	the watershed. The main sources of Ca^{2+} and SO_4^{2-} are likely to be from the dissolution of gypsum
235	and anhydrite. Thus, there is a positive correlation between $Ca^{2\scriptscriptstyle +}$ and $SO_4{}^{2\scriptscriptstyle -}.$ However, $Ca^{2\scriptscriptstyle +}$
236	concentrations for the groundwater samples are generally lower than $\mathrm{SO}_4{}^{2\text{-}}$ concentration,
237	suggesting that carbonate deposition is taking place (Sun, 2010). Similarly, Na ⁺ and Cl ⁻ come mainly
238	from halite dissolution and/or evaporation. To further confirm quantitative relations between the
239	eight main ion components, correlation coefficients for all variables were calculated using the
240	reference vectors (Table 3). Positive correlations exist between Ca^{2+} and Mg^{2+} ($r = 0.99$), Ca^{2+} and
241	SO_4^{2-} ($r = 0.97$), and Mg^{2+} and SO_4^{2-} ($r = 0.95$), respectively. Moreover, there is a high correlation
242	between Na ⁺ and Cl ⁻ ($r = 0.96$). These results are consistent with the SOM maps.

The variation of DBI from minimum to maximum cluster is shown in Fig. 6. The minimum DBI was obtained for a number of groups equal to 4. After determining the optimal number of clusters, the HCA with Ward's method was implemented. Figure 7 shows the hierarchical cluster dendrogram for the SOM nodes. According to the figure, 24 SOM nodes were divided into four different groups.

The pattern classification maps for the four groups are shown in Fig. 8. All samples were classified into these groups (nodes). Group 2 (lower right part) of the pattern classification map is linked to high concentration of Na⁺, Cl⁻, and $SO_4^{2^-}$. These features are also observed for the same location in respective component plane (Fig. 5). On the other hand, all ions (except for HCO₃⁻) in Group 1 located at the upper left part have lower concentrations than corresponding ones for the other groups (Fig. 5).

The groundwater chemistry characteristics for respective groups using standardized reference vectors are shown in Fig. 9. Group 1 is characterized by relatively high concentration of HCO₃⁻. Group 2 displays high concentrations for all cations, SO₄²⁻ and Cl⁻. Group 3 is characterized by high Na⁺ and Cl⁻ concentrations. Group 4 is identified by high Na⁺, K⁺, Cl⁻, HCO₃⁻, and CO₃²⁻ concentrations. In order to further understand group characteristics, average ion concentrations for each group were calculated using raw data (Table 4). The results of this analysis showed that Group 1 can be identified by relatively high contents of SO_4^{2-} and HCO_3^{-} . Other ions have lower concentrations. On the other hand, Groups 2, 3, and 4 are characterized by high concentrations of Cl⁻ and SO₄²⁻. Compared to Chinese drinking water standards, concentrations of Cl⁻ and SO₄²⁻ for Groups 2, 3, and 4 greatly exceed these standards (250 mg/L), respectively. Besides, SOM classification results are consistent with the HCA category from Groups 1, 2, 3, and 4 to Clusters

A, B, C, and D, respectively.

267 Conclusions

In this study, to improve the understanding of hydrogeochemical evolution characteristics and assessing water quality, groundwater samples were collected in the lower part of the Dosit River in the Ordos Basin. By using a Piper trilinear diagram, PCA-HCA and SOM methods, the main conclusions were obtained as follows: (1) TDS for 80% of samples exceeded the Chinese drinking water standard (TDS<1,000 mg/L). The hydrogeochemical facies included Na-HCO₃, Na-SO₄ and Na-Cl types. Anions evolve from HCO₃⁻, SO₄²⁻, and Cl⁻ along the groundwater flow. This process is in accordance with anion evolution characteristics from recharge to discharge areas for aquifers in large sedimentary basins. (2) By applying PCA, three main influencing factors were identified. These correspond to gypsum, halite as well as carbonate dissolution, evaporation, and carbonate precipitation. (3) Using PCA and HCA, water samples were divided into four clusters. In addition, evolution features among these clusters could be identified. (4) Using SOM, we obtained component maps for each variable, which is a readily understandable and visualized map for the strong relationships between Ca^{2+} - Mg²⁺ (r = 0.99), Ca^{2+} - SO₄²⁻ (r = 0.97), Mg²⁺ - SO₄²⁻ (r = 0.95), and Na⁺ - Cl⁻ (r = 0.96). (5) SOM classification results are consistent with the HCA category from Groups 1, 2, 3, and 4 to Clusters A, B, C, and D, respectively.

284 References

Aiuppa A, Bellomo S, Brusca L, D'Alessandro W, Federico C (2003) Natural and anthropogenic
factors affecting groundwater quality of an active volcano (Mt. Etna, Italy). Applied

1	287	geochemistry 18(6): 863-88
3 4	288	Banoeng-Yakubo B, Yidana SM,
5 6 7	289	multivariate statistical me
8 9 10	290	Engineering 13(1): 55-63. h
11 12	291	Brown C (1998) Applied Multivaria
13 14 15	292	Berlin
16 17 18	293	Cloutier V, Lefebvre R, Therrien
19 20 21	294	geochemical data as indica
21 22 23	295	sedimentary rock aquifer sy
24 25 26	296	https://doi.org/10.1016/j.jhg
27 28 29	297	Choi BY, Yun ST, Kim KH, Kim J
30 31	298	of South Korean groundw
32 33 34	299	Geochemical Exploration 1
35 36 37	300	Chebotarev II (1955) Metamorphisr
38 39	301	Cosmochimica Acta 8(1-2):
40 41 42	302	Güler C, Thyne GD, McCray JE,
43 44 45	303	statistical methods for class
46 47	304	455-474 https://doi.org/10
48 49 50	305	García HI González IM (2004) S
51 52	206	monitoring Engineering
53 54 55	300	https://doi.org/10.1016/j.org
56 57 58	307	https://doi.org/10.1016/j.en
59 60	308	Hentati A, Kawamura A, Amaguchi
61 62 63		
64 65		

- Nti E (2009) Hydrochemical analysis of groundwater using ethods -the Volta region, Ghana. KSCE Journal of Civil https://doi.org/10.1007/s12205-009-0055-2
 - ate Statistics in Geohydrology and Related Sciences. Springer,
- R, Savard MM (2008) Multivariate statistical analysis of ative of the hydrogeochemical evolution of groundwater in a vstem. Journal of Hydrology 353(3-4): 294-313. vdrol.2008.02.015
- W, Kim HM, Koh YK (2014) Hydrogeochmical interpretation
- rater monitoring data using self-organizing maps. Journal of 37: 73-84. https://doi.org/10.1016/j.gexplo.2013.12.001
- m of natural waters in the crust of weathering-1. Geochimica et
- : 22-48. https://doi.org/10.1016/0016-7037(55)90015-6
- Turner KA (2002) Evaluation of graphical and multivariate ification of water chemistry data. Hydrogeology Journal 10(4):
 - 1007/s10040-002-0196-6
 - Self-organizing map and clustering for wastewater treatment Applications of Artificial Intelligence 17(3): 215-225. gappai.2004.03.004
 - H, Iseri Y (2010) Evaluation of sedimentation vulnerability at

309	small hillside reservoirs in the semi-arid region of Tunisia using the Self-Organizing Map.
310	Geomorphology 122(1-2): 56-64. https://doi.org/10.1016/j.geomorph.2010.05.013
311	Jiang XW, Wan L, Ge S, Cao GL, Hou GC, Hu FS, Wang XS, Li H, Liang SH (2012) A quantitative
312	study on accumulation of age mass around stagnation points in nested flow systems. Water
313	Resource Research 48(12): W12502. https://doi.org/10.1029/2012WR012509
314	Jiang XW, Wan L, Wang JZ, Yin BX, Fu WX, Lin CH (2014) Field identification of groundwater
315	flow systems and hydraulic traps in drainage basins using a geophysical method.
316	Geophysical Research Letter 41(8): 2812-2819. https://doi.org/10.1002/2014GL059579
317	Jiang XW, Wan L, Wang XS, Wang D, Wang H, Wang JZ, Zhang H, Zhang ZY, Zhao KY (2018)
318	A multi-method study of regional groundwater circulation in the Ordos Plateau, NW China.
319	Hydrogeology Journal. https://doi.org/10.1007/s10040-018-1731-4
320	Jin YH, Kawamura A, Park SC, Nakagawa N, Amaguchi H, Olsson J (2011) Spatiotemporal
321	classification of environmental monitoring data in the Yeongsan River basin, Korea, using
322	self-organizing maps. Journal of Environmental Monitoring 13(10): 2886-2894.
323	https://doi.org/10.1039/C1EM10132C
324	Kalteh AM, Berndtsson R (2007) Interpolating monthly precipitation by self-organizing map
325	(SOM) and multilayer perceptron (MLP). Hydrological Sciences Journal 52(2):305-317.
326	https://doi.org/10.1623/hysj.52.2.305
327	Kohonen T (2001) Self-organizing maps, third ed. Springer, Berlin
328	Kohonen T (1982) Self-organized formation of topologically correct feature maps. Biological
329	Cybernetics 43(1): 59-69. https://doi.org/10.1007/BF00337288
330	Morell L, Giménez E, Esteller MV (1996) Application of principal component analysis to the study
	15

of salinization on the Castellon Plain (Spain). Science of the Total Environment 177(1-3):

332 161-171. https://doi.org/10.1016/0048-9697(95)04893-6

Montcoudiol N, Molson J, Lemieux JM (2015) Groundwater geochemistry of the Outaouais Region
(Québec Canada): a regional-scale study. Hydrogeololgy Journal 23(2): 377-396.
https://doi.org/10.1007/s10040-014-1190-5

- Nakagawa K, Amano H, Asakura H, Berndtsson R (2016) Spatial trends of nitrate pollution and
 groundwater chemistry in Shimabara, Nagasaki, Japan. Environmental Earth Sciences 75:
 234. https://doi.org/10.1007/s12665-015-4971-9
- Nakagawa K, Amano H, Kawamura A, Berndtsson R (2017) Classification of groundwater
 chemistry in Shimabara, using self-organizing maps. Hydrology Research 48(3): 840-850.
 https://doi.org/10.2166/nh.2016.072
- Nguyen TT, Kawamura A, Tong TN, Nakagawa N, Amaguchi H, Gilbuena R (2015) Clustering
 spatio-seasonal hydrogeochemical data using self-organizing maps for groundwater
 quality assessment in the Red River Delta, Vietnam. Journal of Hydrology 522: 661-673.
 https://doi.org/10.1016/j.jhydrol.2015.01.023
- Nishiyama K, Endo S, Jinno K, Betracchi Uvo C, Olsson J, Berndtsson R (2007) Identification of
 typical synoptic patterns causing heavy rainfall in the rainy season in Japan by a SelfOrganizing Map. Atmospheric Research 83 (2-4): 185-200.
- 349 https://doi.org/10.1016/j.atmosres.2005.10.015
 - Parkhurst DL, Appelo CAJ (2013) Description of input and examples for PHREEQC version 3-A
 computer program for speciation, batch-reaction, one-dimensional transport, and inverse
 geochemical calculation. U.S. Geological Survey Techniques and Methods, book 6, chap.

353	A43. http://pubs.usgs.gov/tm/06/a43. Accessed 8 July 2018											
354	Piper AM (1944) A graphic procedure in the geo chemical interpretation of water analysis. Trans											
355	Am Geophys Union 25(6): 914-928. https://doi.org/10.1029/TR025i006p00914											
356	Schoeller H (1972) Geochemistry of ground water. In: Brown RH (ed) Ground-water studies: an											
357	international guide for research and practice. UNESCO, Paris, pp 457-498											
358	Singhal BBS, Gupta RP (2010) Applied hydrogeology of fractured rocks, Second Edition. Springer											
359	Netherland, Dordrecht											
360	Sun FQ (2010) Research on groundwater circulation and environment effect of Dosit River in Ordos											
361	Basin. Changan University (in Chinese with English abstract)											
362	Vesanto J, Himberg J, Alhoniemi E, Parahankangas J (2000) SOM Toolbox for Matlab 5, Helsinki											
363	University of Technology Report A57											
364	Wang H, Jiang XW, Wan L, Han G, Guo H (2015) Hydrogeochemical characterization of											
365	groundwater flow systems in the discharge areas of a river basin. Journal of Hydrology											
366	527: 433-441. https://doi.org/10.1016/j.jhydrol.2015.04.063											
367	Wang JZ, Wörman A, Bresciani E, Wan L, Wang XS, Jiang XW (2016) On the use of late-time											
368	peaks of residence time distributions for the characterization of hierarchically nested											
369	groundwater flow systems. Journal of Hydrology 543: 47-58.											
370	https://doi.org/10.1016/j.jhydrol.2016.04.034											
371	Yin L, Hou GC, Su XS, Wang D, Dong J, Hao Y, Wang X (2011) Isotopes (δD and $\delta^{18}O$) in											
372	precipitation, groundwater and surface water in the Ordos Plateau, China: implications with											
373	respect to groundwater recharge and circulation. Hydrogeology Journal 19(2): 429-443.											
374	https://doi.org/10.1007/s10040-010-0671-4											
	17											
:												

1	375	Yu H, Song Y, Liu R, Pan H, Xiang L, Qian F (2014) Identifying changes in dissolved organic
2		
3	376	matter content and characteristics by fluorescence spectroscopy coupled with self-
4		5 1 15 1
5	277	
7	311	organizing map and classification and regression tree analysis during wastewater treatment.
8		
9	378	Chemosphere 113: 79-86. https://doi.org/10.1016/j.chemosphere.2014.04.020
10		
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12	519	
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2 3 4	398	Figure	Captions
5 6 7	399	Fig. 1	Topography and groundwater sample points for the research area
8 9 0	400	Fig. 2	Piper trilinear diagram for 20 samples
1 2 3	401	Fig. 3	Scatter plot for two principal components with respect to clusters (a) Factors 1 and 2; (b)
4 5	402		Factors 2 and 3
6 7 8	403	Fig. 4	Spatial distribution of respective cluster in the study area
9 0 1	404	Fig. 5	Component planes for (Na ⁺ , K ⁺ , Ca ²⁺ , Mg ²⁺ , Cl ⁻ , SO ₄ ²⁻ , HCO ₃ ⁻ , and CO ₃ ²⁻)
2 3 4	405	Fig. 6	Variation of DBI values with optimal number of groups
5	406	Fig. 7	Dendrogram for respective group using node numbers
7 8 9	407	Fig. 8	Pattern classification map of the four groups by the SOM
0 1 2	408	Fig. 9	Radar charts for respective groups using standardized data
3 4 5	409		
6 7	410	Table c	captions
8 9 0	411	Table 1	Statistical summary of water samples (eight parameters, pH, TDS)
1 2 3	412	Table 2	Results of principal component analysis
4 5 6	413	Table 3	Correlation coefficients among eight physicochemical variables using standardized data
7 8	414	Table 4	Mean of the eight variables for four groups and all data
9 0 1			
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0 1 2			19
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Figure 1



Figure 2













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9	20	8	17
10	21	13	18
15	22		23
			24





	Na ⁺	\mathbf{K}^+	Ca ²⁺	Mg^{2+}	Cl-	SO4 ²⁻	HCO ₃ -	CO ₃ ²⁻	TDS	pH
	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	
Maximum	812.8	7.0	235.0	154.0	813.3	2072.3	243.5	12.6	3973.9	9.0
Minimum	156.7	0.9	3.0	0.2	65.4	100.4	73.4	0.0	454.7	8.0
Mean	543.7	3.8	56.5	36.4	466.8	698.8	157.1	4.7	1889.3	8.4
SD	200.6	2.0	67.2	48.9	226.1	491.2	48.8	6.1	913.6	0.3

Table 1 Statistical summary of water samples (eight parameters, pH, TDS)

SD means standard deviation.

n = 20

	Components				
	Factor 1	Factor 2	Factor 3		
Na ⁺	0.80	-0.14	-0.51		
K^+	0.81	0.53	-0.05		
Ca^{2+}	0.86	-0.01	0.50		
Mg^{2+}	0.85	0.06	0.49		
Cl-	0.83	0.02	-0.52		
SO 4 ²⁻	0.94	-0.1	0.25		
HCO ₃ -	-0.53	0.66	0.35		
CO3 ²⁻	0.07	0.93	-0.28		
Eigen values	4.62	1.61	1.29		
% of variance	57.8	20.1	16.1		
Cumulative %	57.8	77.9	94.0		

 Table 2 Results of principal component analysis

	K^+	Ca^{2+}	Mg^{2+}	Cl-	SO ₄ ²⁻	HCO ₃ -	CO3 ²⁻
Na ⁺	0.76**	0.72**	0.7^{**}	0.96**	0.85**	-0.76**	0.03
\mathbf{K}^+		0.78^{**}	0.81^{**}	0.86**	0.79**	-0.34	0.54**
Ca ²⁺			0.99**	0.74^{**}	0.97**	-0.66**	-0.03
Mg^{2+}				0.76^{**}	0.95**	-0.63**	0.05
Cl-					0.83**	-0.68**	0.23
SO ₄ ²⁻						-0.75**	-0.07
HCO ₃ -							0.51^{*}

Table 3 Correlation coefficients among 8 physicochemical parameters

* Correlations significant at p = 0.05

** Correlations significant at p = 0.01

n = 20

	Na ⁺	\mathbf{K}^+	Ca^{2+}	Mg^{2+}	Cl-	SO_4^{2-}	HCO ₃ -	CO ₃ ²⁻
	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L	mg/L
Group 1	184.9	1.2	9.5	3.7	75.8	154.2	200.9	2.2
Group 2	670.5	6.5	207.7	146.6	653.6	1664.6	125.3	4.2
Group 3	609.6	2.8	33.7	17.6	500.9	634.6	122.4	0
Group 4	571.8	5.1	34.7	22.7	523.3	573.9	197.2	12.6
Whole data	543.7	3.8	56.5	36.4	466.8	698.8	157.1	4.8

 Table 4 Mean values of 8 parameters for 4 groups and whole data