1	Rapid counting method for 2-MIB-producing cyanobacteria
2	(Pseudanabaena sp.) using fluorescence detection of phycocyanin
3	pigments in algal cells
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11 Abstract

12 The sudden occurrence of odor in drinking water is one of the significant challenges for 13 drinking water utilities. The concentrations of 2-methylisoborneol (2-MIB)-producing 14 cyanobacteria, Pseudanabaena sp., can be a surrogate indicator for alarming the potential of 15 odor occurrence. This study aimed to develop a rapid and automatic method to accurately 16 measure Pseudanabaena sp. concentrations using fluorescence detection of auto-fluorescent 17 chemical in cyanobacteria, phycocyanin. This study first identified that increases in 2-MIB 18 concentrations in a lake almost simultaneously occurred according to the rise in the counts of 19 *Pseudanabaena* sp., which was only <0.5% of the entire algae population. The developed 20 method using a phycocyanin filter and threshold settings in fluorescence microscopy (algal cell 21 width and length, fluorescence intensity, and shape) confirmed that almost exclusively 22 cyanobacteria appeared in fluorescent photos, unlike another commonly-used method using a 23 chlorophyll filter that captured all algae. As a result, a high correlation ($R^2=0.998$) was 24 observed between manually-counted and automatically-counted concentrations of 25 Pseudanabaena sp. The detection limit of this method was determined at 11 units/mL. Further, 26 the new method's accuracy using a phycocyanin filter was 88%, while the previously 27 developed method using a chlorophyll filter was approximately 62%. These results suggest that 28 the developed method with a phycocyanin filter can provide a more reliable warning of 29 potential 2-MIB occurrence in drinking water sources.

30 Keywords: algae counting; 2-MIB; cyanobacteria; phycocyanin; drinking water;
31 *Pseudanabaena* sp.

33 Introduction

34 The occurrence of odor in drinking water is a growing concern worldwide. The main odorous compounds from algae in drinking water sources are geosmin and 2-methylisoborneol 35 (2- MIB). These compounds are secondary metabolites mainly originating from cyanobacteria, 36 and the human detection limit of those two compounds is low (<10 ng/L)¹⁻³. Musty odor and 37 38 earthy off-flavors caused by geosmin or 2-MIB in drinking water are significant causes of consumer complaints^{4, 5}. Various cyanobacteria such as Aphanizomenon, Pseudanabaena, 39 Planktothrix can produce 2-MIB and cause odor and taste problems ⁶⁻⁸. Among the numerous 40 41 2-MIB-producing cyanobacteria, Pseudanabaena sp. has been notably identified as a significant odor-causing alga in rivers and lakes globally⁹ and particularly in Asia¹⁰⁻¹². In 2019, 42 Pseudanabaena sp. was detected in Lake Y. (Japan) and identified as responsible for the 2-43 MIB odor events ¹³. The production of 2-MIB by *Pseudanabaena* sp. can vary from 14 to 147 44 fg/cell depending on various factors (e.g. strain, light, and growth medium)^{14, 15}. Although the 45 pre-treatment process using powdered activated carbon (PAC) can remove 2-MIB 46 concentrations ^{16, 17}, the start of PAC doses can be frequently delayed due to the difficulty of 47 48 determining odor concentrations on-site at drinking water treatment plants, resulting in periods 49 of non-compliance with the 2-MIB regulations (e.g., 10 ng/L in Japan).

The concentrations of odorous compounds are typically determined using gas chromatography-mass spectrometry (GC/MS). Because GC/MS is costly and requires a highskilled technician, 2-MIB analysis is only regularly performed (e.g., once a week), even during the odor-occurring period. Any fast detection of odor occurrence can minimize the level and frequency of non-compliance. Recently developed techniques such as artificial olfactory or gustatory systems are gaining interest for the analysis of 2-MIB: a bioelectronic nose can detect 2-MIB in water without any pretreatment ¹⁸, and a potentiometric E-tongue system can analyze 2-MIB concentrations in drinkable water ¹⁹. However, the capacity of these methods for detecting 2-MIB concentrations below 10 ng/L has yet to be established. Quantitative real-time polymerase chain reaction (q-PCR) is also a promising tool for 2-MIB monitoring. The presence and quantity of genes necessary to produce 2-MIB can be analyzed in the water using specific primers ¹⁵. However, when 2-MIB concentration is below 5 ng/L, the genes coding for 2-MIB synthetase cannot be detected. As a result, the rapid analysis of 2-MIB concentrations in surface water (<10 ng/L) remains a challenge.

64 Monitoring the concentrations of 2-MIB-producing algae is an alternative method for 65 predicting odor events. Odor-producing cyanobacteria are commonly measured by microscopic 66 enumeration in bright-field observation. However, it is time-consuming and requires trained 67 personnel. Although digital flow cytometry can be applied to enhance the speed without trained personnel ²⁰, the channel of the flow cytometer can be readily clogged. Moreover, it has been 68 69 mainly used to monitor large marine algae; and despite recent progress, data on freshwater cyanobacteria are limited ²¹ ^{22, 23}. Although a recent study ²⁴ aimed to sort river water 70 71 cyanobacteria into major genera, including Pseudanabaena, the classification was only semi-72 automated due to the requirement of manually checking the photos to identify wrongly counted 73 or uncounted particles. As a result, automated counting of only odor-causing algae in surface 74 waters by digital microscopy has yet to be successful.

A recent study by the authors ¹³ developed an automated counting technique for 2-MIBproducing *Pseudanabaena* sp. Firstly, algae photos are captured, and each alga in the pictures is analyzed based on auto-fluorescence emitted from chlorophyll. After separating cyanobacteria from other types of algae based on their high fluorescence intensity, the unique dimensions (length and width) of *Pseudanabaena* sp. are utilized to enumerate their concentrations. However, long and narrow algae (such as Ankistrodesmus and Nitzschia), often present in high concentrations in lake waters, cause an overestimation of *Pseudanabaena* sp.
concentrations ¹³. The method's accuracy can be improved by eliminating the impact of algae
other than cyanobacteria. This study suggests an alternative approach that utilizes phycocyanin,
a pigment specific to cyanobacteria, instead of chlorophyll a and b. Phycocyanin is commonly
used to detect cyanobacteria ²⁵. The phycocyanin concentration in *Pseudanabaena* sp. cells
varies depending on photoperiod, light intensity, and growth states ^{14, 26, 27}.

This study aims to develop a high-precision and rapid analytical protocol for automatically measuring *Pseudanabaena* sp. concentrations using phycocyanin in algal cells. The accuracy of the developed method was evaluated using the manually-counted algal concentrations in the bright field and the auto-counted algal concentrations using fluorescence emitted by (a) phycocyanin or (b) chlorophyll a and b. The developed method can enhance the reliability of automatically counting *Pseudanabaena* sp. to predict 2-MIB occurrence in drinking water sources.

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95 Materials and Methods

96 Water samples

97 The occurrence of 2-MIB indicated by an increase in *Pseudanabaena* sp. concentrations was 98 evaluated in a lake in Nagasaki Prefecture, Japan. The lake (denoted as Lake Y) is used as a 99 drinking water source. Water samples were collected weekly at a lake water intake, from which 100 the lake water was transported to a drinking water treatment plant. The samples were collected 101 from mid-March to July 2022. It should be noted that the weekly monitoring of 2-MIB and 102 *Pseudanabaena* sp. concentrations continued until the end of September 2022. However, 103 *Pseudanabaena* sp. was absent in the samples from July, and there was no increase in 2-MIB 104 concentration. It was consistent with the trend of odor occurrence in Lake Y; 2-MIB peaks are 105 detected in spring from March to June over the last four years. The water quality is available 106 in **Table S1**. In addition, river water (denoted as River U) in Nagasaki Prefecture, Japan, was 107 also used for verification. The collected water samples were stored at 4 °C and analyzed within 108 four days after collection.

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110 **Odor quantification**

The samples collected from Lake Y were analyzed to detect geosmin and 2-MIB ¹³. In brief, a purge and trap concentrator (PT7000, GL Science, Tokyo, Japan) was used to extract the two compounds in each sample. Then, concentrations of geosmin and 2-MIB were determined using a gas chromatograph–mass spectrometer (GCMS-QP-2010Plus, Shimazu, Kyoto, Japan).

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117 Microscope and sample preparation

118 Lake Y water samples were analyzed weekly to determine algae concentrations. 119 Observations were conducted using an all-in-one Fluorescence Microscope (BZ-X800, 120 Keyence Co.; Osaka, Japan) with objective lenses (Nikon, Japan). A sample volume of $300 \,\mu\text{L}$ 121 of water was inserted into a plankton counting polycarbonate slide (MPC-200, Matsunami 122 Glass Industry; Osaka, Japan). The microscope slide has a specimen capacity of 0.1 mL and a 123 chamber depth of 1 mm. A resting time of 10 min is sufficient to allow the algae to settle at the 124 bottom of the slide. 125 Firstly, the bright-field observation was conducted to identify the accurate concentration 126 of each alga. The natural counting unit (single cell, filament, or colony) determined the number 127 of algae in 0.1 mL. Pseudanabaena sp. and Pseudanabaena limnetica were separately counted 128 in the bright-field observation. Secondly, fluorescence observations were performed using two 129 filters targeting auto-fluorescence emitted by chlorophyll and phycocyanin. The filter targeting 130 chlorophyll a and b (Chroma Technology Japan; Yokohama, Japan) has excitation and 131 emission wavelengths of 440 ± 20 nm and ≥ 590 nm, respectively. In addition, another filter 132 with excitation and emission wavelengths of 592 ± 11 nm and 630 ± 15 nm, respectively 133 (Chroma Technology, Yokohama, Japan) was used to identify phycocyanin fluorescence in algal cells (Fig. S1)²⁸. 134

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136 Auto-counting method development and validation

137 Counting algal concentrations using the photos captured under fluorescence observations were performed similarly to a previous study ¹³. Firstly, 12 pictures of the sample were captured 138 139 using the phycocyanin filter. Each image has a total area of 9.85 mm², equal to a sample volume 140 of 118 µL. Then using the chlorophyll filter, the same area was divided into 49 sequential 141 photos captured in auto-focus and auto-moving stage modes. Secondly, using hybrid cell count 142 software (BZ-H3C, Keyence Co.; Osaka, Japan), the characteristics of each alga were analyzed, 143 and they were sorted based on their dimensions (e.g., width and length), fluorescence intensity, 144 and shape (circularity). The counting software utilizes the minimum bounding box method to 145 determine length and width. In addition, the software calculated circularity as 4π multiplied by 146 object area divided by object perimeter. Finally, the software extracted the target algal cells 147 with pre-set object properties, and they were recognized as the targeted 2-MIB-producing 148 cyanobacteria. Analysis of each sample was performed in duplicate. The image capture and

149 software analysis were completed within 5 min (excluding the resting time) using phycocyanin 150 fluorescence and within 1 h for chlorophyll fluorescence. The auto-counting method was 151 evaluated by correlating manually and automatically counted algal concentrations in Lake Y 152 and River U samples.

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155 **Results and Discussion**

156 Variation of 2-MIB concentrations

157 Weekly analysis of algal concentrations in Lake Y found that increases in 158 Pseudanabaena sp. and 2-MIB concentrations occurred concurrently from mid-March (Fig. 159 1A). On March 25, Pseudanabaena sp. concentrations were identified at 370 units/mL, and 2-160 MIB concentrations were also high at 7.4 ng/L. At the maximum concentration of 370 units/mL, 161 Pseudanabaena sp. accounted for only 0.49% of the entire algal community (Fig. S2). No other 162 2-MIB-producing cyanobacteria were identified. This is consistent with a previous study ¹³ 163 where an odor event was attributed to the minor population of 2-MIB-producing 164 Pseudanabaena sp. Due to the increased 2-MIB concentrations identified on March 25, the 165 water treatment plant dosed copper sulfate, an algicide, into the lake on March 29, and the 166 concentration of *Pseudanabaena* sp. rapidly decreased to 9 units/mL after ten days (April 8) (Fig. 1A). However, on April 1 (three days after copper sulfate treatment), 2-MIB 167 168 concentration reached as high as 13.4 ng/L (Fig. 1A). Because copper (Cu) can cause damages on algal cells, toxin or odor compounds can be released from the algal cells ²⁹⁻³¹. Therefore, it 169 170 can be assumed that the intracellular bound 2-MIB was released in the water after damaging

- *Pseudanabaena* sp. cells ^{32, 33}. Following the reduction of *Pseudanabaena* sp. concentrations,
- the 2-MIB concentration also decreased and stabilized at approximately 3 ng/L (Fig. 1A).



Fig. 1 Concentration of (A) 2-MIB, geosmin, and *Pseudanabaena* sp., and (B) all algae in
Lake Y from March 11 to July 1, 2022. Algal concentrations were quantified via bright-field
observations. Vertical grey lines represent copper sulfate treatment.

179 From April 15, increases in Pseudanabaena sp. and 2-MIB concentrations occurred 180 concurrently (Fig. 1A). On May 13, Pseudanabaena sp. concentrations were identified at 50 181 units/mL, and 2-MIB concentrations reached 4.1 ng/L. As the water temperature rose in Lake 182 Y from 16.0 to 19.3 °C from May 13 to May 20 (Table S1), another copper sulfate treatment to the lake was conducted on May 20. On May 27, one week after treatment, concentrations of 183 184 Pseudanabaena sp. decreased to 8 units/mL. and 2-MIB to 0.7 ng/L. Cyanobacteria concentrations decreased slightly from 3,000 to 2,800 units/mL (Table S2). However, the 185 186 diminution of *Pseudanabaena* sp. and cyanobacteria concentrations had occurred one week 187 before the copper sulfate treatment; thus, the decrease cannot be solely attributed to the 188 algicidal treatment. From June, heavy rainfalls occurred (Fig. S3), and the high precipitations 189 likely flushed the lake water containing 2-MIB and Pseudanabaena sp. After that, 2-MIB 190 concentration was relatively stable below 1.0 ng/L, and Pseudanabaena sp. concentrations 191 remained below 23 units/mL (Fig. 1A). The copper sulfate treatment can mitigate the 192 proliferation of cyanobacteria, including 2-MIB-producing *Pseudanabaena* sp., avoiding a 193 further increase in their population.

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5 The potential of Pseudanabaena sp. to predict the 2-MIB occurrence

This study identified that 2-MIB and *Pseudanabaena* sp. concentrations were somewhat correlated with an R² of 0.673 (r = 0.820, p = .00377) (Fig. 2). In contrast, no correlation was found for the other odor-producing algae in Lake Y, such as *Pseudanabaena limnetica* and *Staurastum* sp. with R² of 0.042 and 0.015, respectively (Fig. S4). Furthermore, no other 2-MIB-producing cyanobacteria, such as Aphanizomenon or Planktothrix, were identified during the study. This indicates that *Pseudanabaena* sp. is likely responsible for the odor event in Lake Y. Although the regulatory limit for 2-MIB in drinking water in Japan is set at 10 ng/L, most water treatment plants maintain 2-MIB concentrations below 3 ng/L as a
conservative measure to avoid customer complaints. In fact, during the odor event, a drinking
water treatment plant managing Lake Y was mainly operated using the other drinking water
sources not impacted by the odor event. Therefore, increasing *Pseudanabaena* sp.
concentrations in lake water could indicate a 2-MIB concentrations increase and potential odor
event occurrence.

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Fig. 2 Correlation between concentrations of 2-MIB and *Pseudanabaena* sp. in Lake Y from March 18 to July 1, 2022. The samples collected after the copper sulfate treatments (April 1 and 8 and May 27) were not included because the copper sulfate treatment impacted the algal concentrations.

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216 Advantages of using a phycocyanin filter in fluorescence observation

217 During the assessment from March 18 to July 1, photos of algae containing 218 autofluorescent pigments (e.g., chlorophyll and phycocyanin) were captured using filters

targeting chlorophyll or phycocyanin fluorescence. Phycocyanin is a pigment present only in 219 220 cyanobacteria, including *Pseudanabaena* sp., red algae (Rhodophytes), and Cryptophytes such as Cryptomonas ³⁴. However, red algae are relatively rare in freshwater ³⁵, and Cryptophytes 221 are non-filamentous ³⁶. Thus, the filamentous cyanobacteria (e.g., *Pseudanabaena* sp.) and 222 223 Cryptomonas were separated based on shape. Green algae and diatoms contain only 224 chlorophyll, whereas cyanobacteria contain chlorophyll and phycocyanin. As a result, green 225 algae (e.g., Scenedesmus) and diatoms (e.g., Nitzschia) were not visible in the photos taken 226 using the phycocyanin filter (Fig. 3). In some cases, when the filament was trapped in 227 aggregates with other algae or particles, the obtained photos became blurry, leading to the 228 miscounting of Pseudanabaena sp. (i.e., underestimation). In addition, all algae appeared in 229 the pictures using the chlorophyll filter, whereas the phycocyanin filter specifically targeted 230 cyanobacteria.





Fig. 3 Microscopic photos of *Pseudanabaena* sp. and other algae in Lake Y: (A) bright-field,

234 (B) chlorophyll fluorescence, and (C) phycocyanin fluorescence. The white line is 50 μm wide.

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236 The main benefit of using the phycocyanin filter over the chlorophyll filter is the 237 improved accuracy due to the capacity to capture cyanobacteria only. Green algae and diatoms 238 do not appear on the captured photos with the phycocyanin filter, significantly reducing the 239 number of algae to analyze, minimizing counting errors, and overestimating. Further, the 240 accuracy was maintained even when algae with shapes similar to Pseudanabaena sp., such as 241 Nitzschia, were in abundance., For example, samples contained high concentrations of Nitzschia (up to 16.8×10⁴ units/mL) on April 1. With the chlorophyll filter, Nitzschia caused 242 many errors and overestimation of auto-counted concentrations of *Pseudanabaena* sp. (Fig. 5). 243 In fact, the leading causes of miscounts in the previous study ¹³ included the miscounting of 244 245 other green algae or diatoms with similar dimensions, moving plankton, double counting in 246 segmented photos, and other cyanobacteria. In contrast, high concentrations of Nitzschia did 247 not impact accuracy with the phycocyanin filter.

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249 Auto-counting method development

A method for determining *Pseudanabaena* sp. concentrations from the captured photos was established based on the capability of the cell counting software, as described in **Fig. 4**. After image acquisition, each fluorescent object in the captured image was set to be automatically sorted based on its properties (e.g., fluorescence intensity, dimensions, and shape) using the cell counting software. Firstly, all objects on the image with fluorescence intensity above 20 (a.u.) are selected. Then objects with dimensions (length and width) outside 256 the targeted range are removed from the selection. Last, the objects with a circular shape are 257 removed from the selection, keeping the more elongated objects (filaments) in the selection. 258 The final conditions to automatically count *Pseudanabaena* sp. include a length of $\geq 10 \,\mu\text{m}$, a 259 width of 1.5–12 µm, and circularity <0.9 (i.e., not circular objects) (Table S4). Because more 260 suitable images can be attained due to the use of the phycocyanin filter, the developed method 261 in this study allows for the counting of shorter filaments ($\geq 10 \mu m$) in comparison with the 262 previous technique ($\geq 15 \mu m$)¹³. During the study, only two *Pseudanabaena* sp. filaments shorter than 10 µm were observed, indicating the validity of the length setting. The other causes 263 264 of failing to count Pseudanabaena sp. include weak fluorescence intensity or folded form (i.e., 265 wider width). Throughout 16 sampling occasions in Lake Y in the study from March 18 to July 266 1,209 filaments of Pseudanabaena sp. were manually identified and counted in bright 267 observation. In contrast, as high as 184 filaments (88% accuracy) were correctly and 268 automatically calculated based on the fluorescent-based method established using the 269 phycocyanin filter in this study. It is noted that the technique in this study analyzed a sample 270 volume of 0.1 mL within 15 min, including the resting time, which is by far faster than a previous study ¹³ using the chlorophyll filter. Further, the improved technique did not use 271 272 sequential photos for image capturing; thus, there was no error due to double counting.



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Fig. 4 Flowchart describing the technique determining *Pseudanabaena* sp. counts using
fluorescence microscopy, phycocyanin filter, and cell counting software.

278 Accuracy of the phycocyanin-based method

Because of the improved accuracy, the concentrations of automatically-counted 279 280 Pseudanabaena sp. were highly correlated with manually counted Pseudanabaena sp. (Fig. 5). For Lake Y, the R² value of the phycocyanin-based method was 0.997 (r = 0.998, $p = 2.483 \times 10^{-10}$ 281 ¹⁹), while that of the chlorophyll-based method was 0.022 (r = 0.148, p = 0.196). For River U, 282 the R² value of the phycocyanin-based method was 0.809 (r = 0.899, p = 0.147), while that of 283 the chlorophyll-based method was 0.063 (r = 0.263, p = 0.546) (Fig. 6). The abundance of 284 285 Nitzschia was the leading cause of errors and overestimation for the chlorophyll method. The 286 method detection limit (MDL) (11 units/mL) was calculated with 8 samples of 20 units/mL using the following formula: $MDL = Value of the student's t distribution \times standard deviation.$ 287 288 Overall, the phycocyanin-based method developed in this study allows us to analyze a larger 289 volume quickly and provides an accurate analysis of *Pseudanabaena* sp. concentrations.



Fig. 5 Correlation between manual and auto counts of *Pseudanabaena* sp. in Lake Y from
March 18 to July 1, 2022, using the phycocyanin (red) or chlorophyll (blue) filter. A line of
equality (in green) was added for comparison. The sample volume was 100 μL.



Fig. 6 Correlation between manual and auto counts of *Pseudanabaena* sp. in river U from
using the phycocyanin (red) or chlorophyll (blue) filter. A line of equality (in green) was added
for comparison. Sample volume of 100 μL.

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301 Practical implications

The improved counting technique for *Pseudanabaena* sp. in lake waters is an easy and fast way to monitor changes in *Pseudanabaena* sp. concentrations, especially when *Pseudanabaena* sp. is a minority among algae (<1%). Rapid analysis using this technique can allow drinking water treatment plants to take precautionary actions such as increased doses of PAC for odor removal. In addition, fast monitoring of *Pseudanabaena* sp. could also help manage the water source and plan preventive algicide treatment if the concentration of odorproducing *Pseudanabaena* sp. rapidly increases within a week. 309 A limitation of the method developed in this study is the possibility of overestimation in 310 case there is a high concentration of other filamentous cyanobacteria that have a similar size to 311 Pseudanabaena sp. because the criteria used by the software are dimensions and fluorescence 312 intensity only. As *Pseudanabaena* sp. is the dominating filamentous cyanobacteria in Lake Y, 313 further studies would be necessary to adapt the technique to different water sources. Moreover, non-planktonic, benthic cyanobacteria are also known to be responsible for odor events ³⁷. 314 315 However, as those cyanobacteria are periphytic (e.g., attached to submerged surfaces), they are 316 not present in the sampled water. Algae with motility had little to no impact on the method as 317 the exposure time before capture was short (0.33 sec/image); therefore, fixation pretreatment 318 was not used in this study. Another limitation is that the developed method using only 319 fluorescence microscopy photos cannot differentiate between odor-producing cyanobacteria 320 and non-producers. The examples include Pseudanabaena sp. (2-MIB-producing 321 cyanobacteria) and Pseudanabaena limnetica (some strains can produce 2-MIB). Despite the 322 limitations, the technique could detect any increase in filamentous cyanobacteria 323 concentrations, prompting further analysis of the source waters (e.g. odorous compounds or 324 PCR analysis). It is noted that *Pseudanabaena limnetica* was not the major source of 2-MIB peaks in Lake Y during this assessment. In Lake Y, Pseudanabaena limnetica was detected at 325 326 only <10 units/mL. More importantly, *Pseudanabaena limnetica* was not present in the sample during the odor event. Further studies will evaluate the possibility of overestimating 327 328 Pseudanabaena sp. concentrations in water samples due to cyanobacteria with similar 329 dimensions. The width of *Anabaena* sp. (6 to 14.6 µm³⁸) is greater than that of *Pseudanabaena* 330 sp. Therefore, overestimations are unlikely to occur. However, the filament shape of 331 Aphanizomenon sp. and Oscillatoria sp. is similar to that of Pseudanabaena sp.; thus, the 332 overestimations could happen because the software calculates the length and width of the bounding box method. Calculating the filament's average width could help separate 333

Pseudanabaena sp. from similar cyanobacteria such as *Aphanizomenon* sp. and *Oscillatoria*sp. Despite these limitations, the fast analysis developed here can trigger an alarm, allowing
the operators of a drinking water treatment plant to act.

337 It should be noted that the same approaches can be adapted to monitor other algae other 338 than Pseudanabaena sp. For example, some diatoms (e.g., Synedra sp.) are responsible for 339 clogging sand filters in water treatment plants, resulting in more frequent backwashing and reducing drinking water production ³⁹. Concentrations of *Synedra* sp. could be monitored by 340 341 using chlorophyll fluorescence and selecting suitable dimensions and fluorescence intensity. 342 Another application could be monitoring of picophytoplankton concentrations. The presence 343 of picophytoplankton in the water can increase the assimilable organic carbon concentration after disinfection, promoting microbial growth and degrading the water quality ⁴⁰. 344 345 Picophytoplankton concentrations can be automatically determined by integrating their 346 dimensions, chlorophyll, and/or phycocyanin fluorescence intensities into the criteria. 347 Therefore, further research will focus on other problematic algae in drinking water applications.

348

349 **Conclusions**

The increases in 2-MIB concentrations in the studied lake concurrently occurred according
 to the rise in *Pseudanabaena* sp. concentrations, confirming that *Pseudanabaena* sp. can
 be a helpful indicator of odor occurrence.

Automatically measured *Pseudanabaena* sp. concentrations using a phycocyanin filter
 were superior to those using a chlorophyll filter in speed and accuracy.

A high correlation between manually and automatically counted *Pseudanabaena* sp.
 concentrations was observed in the lake water samples; among 209 filaments of

357 *Pseudanabaena* sp. observed during the study, the developed method achieved the 358 identification of as many as 184 filaments with an accuracy of 88%.

• The phycocyanin filter-based technique developed in this study can predict the potential of

- 360 2-MIB occurrence in lakes by rapidly measuring the concentrations of 2-MIB-producing
- 361 cyanobacteria, *Pseudanabaena* sp.

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363 **Conflict of interest**

364 The authors have no conflicts of interest to declare.

365

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369 **References**

- E. C. Wert, J. A. Korak, R. A. Trenholm and F. L. Rosario-Ortiz, *Water Research*, 2014,
 52, 251-259.
- 372 2. S. B. Watson, B. Brownlee, T. Satchwill and E. E. Hargesheimer, *Water Research*, 2000, 34, 2818-2828.
- 374 3. I. H. Suffet, D. Khiari and A. Bruchet, *Water Science and Technology*, 1999, 40, 1-13.
- 375 4. X. Bai, T. Zhang, C. Wang, D. Zong, H. Li and Z. Yang, *Environmental Science and*376 *Pollution Research*, 2017, 24, 2904-2913.
- 377 5. S. Watson, *J Toxicol Environ Health A*, 2004, **67**, 1779-1795.
- M. Su, J. Yu, J. Zhang, H. Chen, W. An, R. D. Vogt, T. Andersen, D. Jia, J. Wang and M. Yang, *Water Res*, 2015, 68, 444-453.

- 380 7. M. Su, M. D. Suruzzaman, Y. Zhu, J. Lu, J. Yu, Y. Zhang and M. Yang, *Journal of Environmental Sciences*, 2021, **110**, 119-128.
- H. M. Franklin, R. Podduturi, N. O. G. Jørgensen, D. T. Roberts, L. Schlüter and M. A.
 Burford, *Chemical Engineering Journal Advances*, 2023, 14, 100455.
- 384 9. C. E. Sotero-Martins A, Santos JAA, et al., *International Journal of Hydrology*, 2021,
 385 5, 214-220.
- J. E. Lee, M. N. Yu, S. Yu and M. Byeon, *Environmental Microbiology Reports*, 2022,
 14, 197-202.
- 388 11. L. Li, S. Yang, S. Yu and Y. Zhang, *Water Research*, 2019, **162**, 180-189.
- 389 12. J.-Y. Jeong, S.-H. Lee, M.-R. Yun, S.-E. Oh, K.-H. Lee and H.-D. Park,
 390 *Microorganisms*, 2021, 9, 2486.
- 391 13. S. Boivin, E. Hasegawa, D. Yamaguchi and T. Fujioka, *Environmental Science: Water* 392 *Research & Technology*, 2021, 7, 1032-1039.
- M. Su, J. Fang, Z. Jia, Y. Su, Y. Zhu, B. Wu, J. C. Little, J. Yu and M. Yang,
 Environmental Research, 2023, 221, 115260.
- 395 15. K.-Y. Lu, Y.-T. Chiu, M. Burch, D. Senoro and T.-F. Lin, *Water Research*, 2019, 164, 114938.
- 16. L. a. W. Ministry of Health, Japan, Overview of water supply services, <u>https://www.mhlw.go.jp/english/policy/health/water_supply/1.html</u>, (accessed 7 July 2022, 2022).
- 400 17. Y. Matsui, Y. Nakano, H. Hiroshi, N. Ando, T. Matsushita and K. Ohno, *Water Science and Technology*, 2010, 62, 2664-2668.
- 402 18. M. Son, D.-g. Cho, J. H. Lim, J. Park, S. Hong, H. J. Ko and T. H. Park, *Biosensors and Bioelectronics*, 2015, 74, 199-206.
- 404 19. L. Lvova, I. Jahatspanian, L. H. C. Mattoso, D. S. Correa, E. Oleneva, A. Legin, C. Di
 405 Natale and R. Paolesse, *Sensors (Basel)*, 2020, 20.
- 406 20. K. Kraft, J. Seppälä, H. Hällfors, S. Suikkanen, P. Ylöstalo, S. Anglès, S. Kielosto, H.
 407 Kuosa, L. Laakso, M. Honkanen, S. Lehtinen, J. Oja and T. Tamminen, *Frontiers in Marine Science*, 2021, 8.
- 409 21. M. D. Graham, J. Cook, J. Graydon, D. Kinniburgh, H. Nelson, S. Pilieci and R. D.
 410 Vinebrooke, *Limnology and Oceanography: Methods*, 2018, 16, 669-679.
- 411 22. L. Barsanti, L. Birindelli and P. Gualtieri, *Environmental Science: Processes & Impacts*,
 412 2021, 23, 1443-1457.
- 413 23. B. M. Owen, C. S. Hallett, J. J. Cosgrove, J. R. Tweedley and N. R. Moheimani,
 414 *Limnology and Oceanography: Methods*, 2022, 20, 400-427.

- 415 24. Y. Mirasbekov, A. Abdimanova, K. Sarkytbayev, K. Samarkhanov, A. Abilkas, D.
 416 Potashnikova, G. Arbuz, Z. Issayev, I. A. Vorobjev, D. V. Malashenkov and N. S.
 417 Barteneva, *Frontiers in Marine Science*, 2021, 8.
- 418 25. H. Almuhtaram, F. A. Kibuye, S. Ajjampur, C. M. Glover, R. Hofmann, V. Gaget, C.
 419 Owen, E. C. Wert and A. Zamyadi, *Ecological Indicators*, 2021, 133, 108442.
- 420 26. Z. Khan, W. O. Wan Maznah, M. S. M. Faradina Merican, P. Convey, N. Najimudin
 421 and S. A. Alias, *Polar Science*, 2019, 20, 3-8.
- 422 27. M. Cegłowska, A. Toruńska-Sitarz, J. Stoń-Egiert, H. Mazur-Marzec and A.
 423 Kosakowska, *Algal Research*, 2020, 47, 101861.
- 424 28. N. Singh, R. Sonani, R. Rastogi and D. Madamvar, 2015.
- 425 29. D. Jančula and B. Maršálek, *Chemosphere*, 2011, **85**, 1415-1422.
- 426 30. H. Xu, J. Brookes, P. Hobson and H. Pei, *Water Research*, 2019, **157**, 64-73.
- 427 31. A. Zamyadi, K. E. Greenstein, C. M. Glover, C. Adams, E. Rosenfeldt and E. C. Wert,
 428 *Water*, 2020, 12, 1105.
- 429 32. L. Li, C. Zhu, C. Xie, C. Shao, S. Yu, L. Zhao and N. Gao, *Water Research*, 2018, 147, 430
 422-428.
- 431 33. H. Li, L. Li, Q. Yin, S. Yu, N. Gao, X. Wang and J. Chen, *Chemical Engineering Journal*, 2022, 440, 135962.
- 433 34. J. Dagnino-Leone, C. P. Figueroa, M. L. Castañeda, A. D. Youlton, A. Vallejos434 Almirall, A. Agurto-Muñoz, J. Pavón Pérez and C. Agurto-Muñoz, *Computational and*435 Structural Biotechnology Journal, 2022, 20, 1506-1527.
- 436 35. F. Nan, J. Feng, J. Lv, Q. Liu, K. Fang, C. Gong and S. Xie, *Scientific Reports*, 2017, 7, 2934.
- 438 36. K. Hoef-Emden and M. Melkonian, *Protist*, 2003, **154**, 371-409.
- 439 37. V. Gaget, H. Almuhtaram, F. Kibuye, P. Hobson, A. Zamyadi, E. Wert and J. D.
 440 Brookes, *Harmful Algae*, 2022, 113, 102185.
- 441 38. R. Li, M. Watanabe and M. M. Watanabe, *Hydrobiologia*, 2000, **438**, 117-138.
- 442 39. H.-B. Jun, Y.-J. Lee, B.-D. Lee and D. R. U. Knappe, *Journal of Water Supply:*443 *Research and Technology-Aqua*, 2001, **50**, 135-148.
- 444 40. T. Okuda, Y. Uehara, T.-Y. Tsai, S. Nakai, M. Akiba, W. Nishijima and M. Okada, 445 *Water Supply*, 2009, **9**, 337-342.