

Don't rely too much on trees:
Evidence from flood mitigation in China

Kaori Tembata^{†§}, Yuki Yamamoto^{‡§*}, Masashi Yamamoto[†],

Ken'ichi Matsumoto[‡]

March 2020

Abstract

Combining a popular flood disaster dataset with climate data and satellite land cover data from China, this paper estimates how forests mitigate the frequency of flooding, resulting in two major findings. First, we confirm that an increase in forest area mitigates the possibility of flood occurrence even after controlling for socioeconomic and meteorological variables and time-invariant individual effects. Second, broadleaf trees and mixed-tree forests have a flood mitigation effect, whereas coniferous trees do not; these results are robust against alternative model specifications. This paper newly corroborates the concept of ecosystem-based disaster risk reduction. While there is an emerging consensus that ecosystems can mitigate natural disasters, there is limited evidence on how ecosystems mitigate disasters. To the best of the authors' knowledge, this study is the first to show that the type of forest is critical for mitigating floods in a rigorous econometric way (survival analysis) spanning numerous areas of interest.

Keywords: flood mitigation, afforestation, forest type, China, remote sensing data, forest ecosystem service

[†]Center for Far Eastern Studies, University of Toyama

[§]Contributed equally

[‡]Graduate School of Fisheries and Environmental Sciences, Nagasaki University

*Corresponding author: Postal address: 1-14 Bunkyo-machi, Nagasaki 852-8521, Japan,
E-mail: y-yamamoto@nagasaki-u.ac.jp

1 Introduction

2 Among all the types of natural disasters occurring worldwide, floods have
3 occurred most frequently over the past couple of decades, accounting for
4 43% of all natural disasters recorded between 1998 and 2017, followed
5 by storms and earthquakes (Wallemacq and House, 2018). During the
6 same period, floods affected approximately two billion people and inflicted
7 economic damage, reaching USD 656 billion. In 2018 alone, 34.2 million
8 people were affected by flooding, and economic losses of USD 19.7 billion
9 were incurred (CRED, 2019). Within the context of disaster risk reduction,
10 the importance of natural ecosystems has gained considerable attention on
11 a global scale. For example, the Millennium Ecosystem Assessment (MEA)
12 emphasizes the use of the natural environment (e.g., mangroves, wetlands,
13 and upland forests) as response options for flood and storm control instead of
14 the physical structures and measures historically employed (e.g., dams and
15 drainage channels) (MEA, 2005).

16 Moreover, the MEA highlights how these ecosystem services are linked to
17 human well-being.¹ Therefore, by impacting environmental security, health,
18 and livelihood, the degradation of ecosystem services negatively affects
19 people's lives. In particular, the loss of forests leads to soil erosion and a

¹Many recent studies found that forest ecosystems could affect rural livelihood (Costanza et al., 2014; Ickowitz et al., 2014; Yamamoto et al., 2019).

20 decrease in the capacity to retain water, thereby increasing the vulnerability
21 of affected people and areas to floods and other natural hazards (Zong and
22 Chen, 2000).

23 Over the years, China has suffered significant flooding. As a
24 countermeasure intended to reduce flood risk, the country has dramatically
25 increased its forest area by introducing the Grain for Green Program (GGP).²
26 The GGP aims to transform steep farmlands into forests to reduce soil
27 erosion and the risks of floods in the upper and middle reaches of the Yellow
28 and Yangtze Rivers, constituting the world's largest payment for ecosystem
29 services. Since the compensation scheme involves local farmers,³ the GGP
30 affects both the natural environment and local livelihood in several ways,
31 including improving the livelihood of farmers (Rodríguez et al., 2016; Wu
32 et al., 2019), protecting ecosystem services and forestland (Xu et al., 2018;
33 Li et al., 2019; Qian et al., 2019; Fan and Xiao, 2020), decreasing water yield
34 (Rodríguez et al., 2016; An et al., 2017; Wang et al., 2019), moderating soil
35 erosion (Lu et al., 2013; Peng et al., 2019; Ye et al., 2019; Wu et al., 2019),
36 and enhancing carbon stock (Song et al., 2015; Peng et al., 2019; Wu et al.,

²While deforestation remains an important issue throughout the world, China increased its forest area from 1.57 million hectares to 2.1 million hectares between 1990 and 2016 (FAO, 2018).

³Each farmer received CNY 300 (USD 43 as of November 2019) per hectare per year and in-kind compensation for 8 years (transformation to ecological forest), 5 years (to economic forest), or 2 years (to grassland) (Delang and Yuan, 2016). Thus, the total compensation payment reached CNY 78.44 billion (USD 11.26 billion) between 2002 and 2005 (Delang and Yuan, 2016).

37 2019).

38 This paper examines the effects of forest cover on flood frequency in
39 China to confirm whether the recent promotion of forest area has contributed
40 to the mitigation of flooding. Specifically, we focus on forest types to
41 examine whether any particular type of forest can help mitigate the risk of
42 floods. While there is an emerging consensus that ecosystems can mitigate
43 natural disasters, there is limited evidence on how ecosystems mitigate flood
44 occurrence. To the best of our knowledge, this study is the first to show
45 that the forest type is critical for mitigating floods in a rigorous econometric
46 way spanning numerous areas of interest. In this study, we applied survival
47 analysis methods to investigate the effects of forest ecosystems on flood
48 occurrence because floods can be assumed to be events occurring with a
49 certain probability during periods. Our analysis also includes socioeconomic
50 and meteorological characteristics as potential confounding factors that most
51 likely affect the occurrence of floods

52 This study contributes to the literature on a debate among hydrological
53 and forestry science on the role of forest ecosystems on flood mitigation.⁴
54 One component of the literature has reported evidence of the effects of
55 deforestation on the occurrence of floods and the corresponding damage

⁴We will discuss the hydrological mechanisms of how forests and floods are related in detail in Section 2.

56 caused by these events. Bradshaw et al. (2007) used cross-country
57 panel data for 56 developing countries from 1990 to 2000 to study the
58 relationship between forest cover and flood frequency. Their statistical
59 analyses demonstrated that the number of flood events was associated
60 with forest-related factors, such as forest cover, natural forest loss, and
61 nonnatural forest cover. By incorporating forest cover attributes into
62 models, their study ultimately found that deforestation caused floods with
63 an increased frequency. The effect of forest cover on flood mitigation is
64 also supported by recent empirical work. Bhattacharjee and Behera (2017,
65 2018) examined whether forest cover can mitigate floods in India. Their
66 investigations revealed that areas with more forest cover were associated with
67 less flood-related damage and highlighted the ability of forests to weaken
68 the adverse impact of climate change incurred by extreme weather events
69 (Bhattacharjee and Behera, 2018). In the study analyzing the impact of
70 public policies on the occurrence of natural disasters in Brazil, Sant'Anna
71 (2018) found that while extreme rainfall increased the frequencies of floods
72 and landslides, negative impacts were mitigated in areas with relatively high
73 forest cover.

74 While the above studies showed that forest cover can have a significant
75 mitigating effect on flood events, others found that this conclusion does not
76 hold (Van Dijk et al., 2009; Ferreira and Ghimire, 2012; Ferreira et al.,

77 2013). In fact, the relationship between forests and floods is a much
78 debated topic insomuch that the roles of forest cover in preventing floods
79 are questioned (CIFOR, 2005). Van Dijk et al. (2009) reanalyzed the work
80 performed by Bradshaw et al. (2007) and argued that the results of the
81 latter are inconclusive when socioeconomic factors are not considered in the
82 estimation; after considering the impact of population density, they found
83 no correlation between forest cover or forest loss and the frequency of floods.
84 The study by Bradshaw et al. (2007) was similarly challenged by Ferreira and
85 Ghimire (2012), who found an insignificant impact of forest cover when the
86 estimation considered other socioeconomic and institutional characteristics.
87 They argued that these factors may be more important than deforestation
88 as determinants of human-induced floods.

89 Indeed, deforestation is not the only way by which humans can impact
90 floods. The consensus in the literature on the economic impacts of natural
91 disasters is that the extent of disaster-related damage is associated with
92 countries' income levels (Kahn, 2005; Noy, 2009; Kellenberg and Mobarak,
93 2008; Ferreira et al., 2013). In addition to income, other socioeconomic
94 factors that most likely affect the frequency of floods and flood-induced
95 damage include a variety of demographic and institutional factors, e.g.,
96 population, urbanization, corruption, and democracy levels (Kahn, 2005;
97 Güneralp et al., 2015; Ferreira and Ghimire, 2012). Furthermore,

98 geographical and meteorological characteristics are considered to be
99 important factors that affect flood occurrence (Zong and Chen, 2000;
100 Sant'Anna, 2018). It is also widely recognized that flood occurrence is
101 affected by land degradation and soil erosion resulting from land use change
102 (Zong and Chen, 2000; Bradshaw et al., 2007). Hence, in addition to forest
103 cover, these factors should be considered when further analyzing the roles
104 forests play in mitigating floods.

105 Moreover, many investigations have linked natural disasters to land use
106 and land cover (Yin and Li, 2001; Van Westen et al., 2008; Van Dijk et al.,
107 2009; Tan-Soo et al., 2016; Wells et al., 2016). To explore these relationships,
108 researchers often apply spatial data to natural hazards and land use and land
109 cover (Bradshaw et al., 2007; Van Dijk et al., 2009; Wells et al., 2016). For
110 instance, Wells et al. (2016) incorporated interview surveys and newspaper
111 articles to spatially analyze whether flood frequency is related to land use
112 in Indonesian Borneo. Their results suggested that the frequency of floods
113 tends to decrease in areas with more logged and intact forests and increase
114 in areas with more extensive oil palm plantations.

115 This study aims to clarify the hypothesis that the existence of forest
116 cover mitigates flood frequency and the mitigation effects differ by forest
117 type. In this sense, our work is also related to ecosystem-based disaster
118 risk reduction (Eco-DRR) or natural-based solutions because forests provide

119 various ecosystem services that reduce hydrological risks, land degradation,
120 and climatic risks (Keesstra et al., 2018; Albert et al., 2019; Calliari et al.,
121 2019; Dorst et al., 2019).

122 **2 The role of forests in water yield**

123 The hydrological impacts of forests have been debated by researchers in the
124 fields of forestry science and hydrology for almost a century (Bruijnzeel,
125 2004). On the one hand, Gentry and Lopez-Parodi (1980) found that the
126 frequency of floods in the Amazon increased due to increased runoff caused
127 by deforestation, although precipitation patterns remained unchanged. On
128 the other hand, Hewlett (1982) observed that the existence of forests did
129 not influence the quantity of water flow. Ultimately, Ferreira et al. (2013)
130 concluded that it was difficult to identify whether forest cover was the
131 sole factor affecting flood occurrence because forest cover changes and
132 socioeconomic conditions both affect the frequency of flooding.

133 More recently, however, it has been acknowledged that the existence
134 of forests or vegetation can contribute to the mitigation of flood risk.
135 Bosch and Hewlett (1982) highlighted that an increase in forest cover can
136 decrease streamflow, while enhanced deforestation leads to an increase in
137 streamflow. Ogden et al. (2013) found that forests reduced the amount of
138 runoff water during the heavy rainy season in Panama, while forests increased

139 the runoff rate during the dry season. Wang et al. (2019) found that forests
140 decreased the water yield in China and attributed this phenomenon to the
141 increased water conservation capacity in afforestation areas. Andréassian
142 (2004) reviewed hydrological studies that conducted experiments with paired
143 watersheds and discovered that deforestation can increase the flood volume
144 and flood peak; in contrast, reforestation is associated with a decreased water
145 yield. Filoso et al. (2017) summarized 308 case studies while focusing on
146 the hydrological impacts of reforestation and mostly found that increasing
147 the extent of forest cover can decrease the water yield. Ellison et al. (2017)
148 revealed that some functions of forests play significant roles in mitigating the
149 occurrence and intensity of floods; for example, forests can disperse water by
150 intercepting and recycling precipitation, promoting upward moisture fluxes,
151 and recharging infiltration and groundwater.

152 In addition, some researchers have discovered that different types of
153 vegetation have varying hydrological effects. Tan-Soo et al. (2016) reported
154 that the conversion of forests into plantations (such as oil palm plantations)
155 led to an increased likelihood of flooding in Malaysia, and Swank and
156 Douglass (1974) observed that the clearing of coniferous forest increased the
157 water yield in the study area more than the clearing of broadleaf forest.
158 However, Brown et al. (2005) noted that the impacts of forest changes on
159 water yield should be quantified based on long-term analyses and found that

160 the effects varied according to the types of vegetation and land use. In this
161 context, Komatsu et al. (2007) demonstrated that broadleaf forest had a
162 greater potential to decrease the water yield in Japan than coniferous forest.

163 Considering the findings of the above literature, the types of vegetation,
164 meteorological conditions, and socioeconomic factors must be considered to
165 investigate the hydrological impacts of forests.

166 **3 Research design**

167 To investigate the relationship between forest cover and flood occurrence in
168 China (focusing particularly on forest types), we employ survival (duration)
169 analysis.⁵ Our analyses are conducted at the subdistrict level from 2001
170 to 2018 considering the availability of relevant data. The flowchart of our
171 estimation procedure is given in Figure 1. In section 3.1, we introduce
172 our dataset, and in section 3.2, we show the empirical framework employed
173 herein. QGIS 2.14.12 and Stata 14.2 were used to conduct the geographical
174 and statistical analyses.

175 **3.1 Data**

176 The forest cover data we employ were obtained from satellite observations
177 provided by Sulla-Menashe et al. (2019). This dataset has been updated

⁵The survival analysis treats time as a continuous variable and can be applied to investigate the repeated and sequential occurrence of events.

178 and is currently available for the period from 2001 to 2018. The dataset
179 comprises global land cover grids with dimensions of 0.05×0.05 degrees based
180 on the International Geosphere-Biosphere Programme (IGBP) classification.
181 In particular, a pixel dominated by woody vegetation (covering over 60% of
182 the pixel) with a tree height higher than 2 m is reported as forest. Based on an
183 identification strategy of observing trees during an annual cycle of leaf-on and
184 leaf-off periods, the dataset provides five forest type classifications: evergreen
185 coniferous, evergreen broadleaf, deciduous coniferous, deciduous broadleaf,
186 and mixed forest.⁶

187 The forest area in China has increased over the last two decades. The
188 broadleaf forest area increased from 4.20 million km^2 in 2001 to 5.04 million
189 km^2 in 2017; the coniferous forest area increased from 0.76 million km^2 to
190 1.22 million km^2 ; and the mixed forest area increased from 15.31 million km^2
191 to 17.77 million km^2 in the same period (Sulla-Menashe et al., 2019).⁷

192 To investigate the effect of each forest type on flood occurrence, we
193 aggregate and recategorize pixels based on broadleaf, coniferous, and mixed
194 forests at the subdistrict level. Figure 2 shows the forest gain by forest
195 type between 2001 and 2017. In particular, broadleaf forest accounts for a
196 large part of the forest gain in northeastern and southern China. Similarly,

⁶Mixed forest consists of a mixture of various forest types.

⁷We aggregate and recategorize the forest type into broadleaf, coniferous, and mixed forest.

197 Figure 3 shows the change in the forest cover rate at the subdistrict level
198 in China between 2001 and 2017. In terms of broadleaf forest, 68.4% of
199 subdistricts experienced forest gain during the study period. Furthermore, a
200 large proportion of subdistricts in northeastern, central, and southern China
201 displayed a gain in forest cover during the study period. However, the forest
202 cover did not change in most of the subdistricts in western China.⁸

203 The flood data were acquired from the Global Active Archive of Large
204 Flood Events, Dartmouth Flood Observatory (Brakenridge, 2012). This
205 dataset has recorded the occurrence of global floods since 1985.⁹ Figure 4
206 shows the number of floods recorded in the database in China between 2001
207 and 2017. Evidently, the number of floods has decreased in China in recent
208 years, whereas the frequency and severity of floods have increased worldwide
209 (Najibi and Devineni, 2018; Wallemacq and House, 2018).

210 The weather data were obtained from the Climate Prediction Center's
211 Global Unified Precipitation dataset provided by the National Oceanic and
212 Atmospheric Administration.¹⁰ This dataset reports global precipitation in

⁸There are few forest areas in the western regions corresponding to the definition that a forest that covers more than 60% of each pixel with a tree height higher than 2 m.

⁹The flood events presented in the Dartmouth Flood Observatory are derived from a variety of news, governmental sources, and remote sensing sources. The dataset provides the flood event data including the location, beginning and ending days, affected areas of flood occurrence as well as the severity of the flood as the indicator of the intensity of the floods. For a more detailed description of the floods in this dataset, see <http://floodobservatory.colorado.edu/index.html>.

¹⁰The data are available at <https://www.esrl.noaa.gov/psd/>.

213 grids of 0.05×0.05 degrees. Our precipitation data refer to the values that are
214 geographically nearest to the center of the corresponding subdistrict. The
215 demographic data were obtained from the National Bureau of Statistics of
216 China.¹¹

217 Table 1 presents the descriptive statistics of our sample.¹² Our dependent
218 variable, *flood*, is a dummy variable that takes a value of one when the flood
219 occurred in the considered subdistrict and zero otherwise, indicating that
220 the probability of flood occurrence is 2.8% for all subdistricts between 2001
221 and 2017. Table 1 also reports the areas of forest cover at the subdistrict
222 level based on the classification of broadleaf, coniferous, and mixed forest.
223 Broadleaf and mixed forest account for a large portion of the observed
224 forest cover, while coniferous forest covers a relatively small area in China.
225 Regarding precipitation, the maximum daily precipitation in a year and the
226 annual average precipitation are also reported in Table 1.

227 **3.2** *Model*

228 We adopt survival analysis with both parametric and semiparametric models
229 to investigate the effects of forest resources on flood occurrence. For the

¹¹See <http://data.stats.gov.cn/english/index.htm>.

¹²We aggregated the dataset to merge the information at the subdistrict level. Detailed information on the data source is summarized in Table S1 of supplemental material.

230 parametric analysis, we use the Weibull hazard function, denoted as

$$h(t|m) = \gamma m t^{m-1}, \quad (1)$$

231 where $\gamma > 0$ and $m > 0$ are parameters. It is common to allow $\gamma = \exp(x'\beta)$
232 to include regressors because this allowance guarantees that $\gamma > 0$. Thus,
233 our hazard function is expressed as

$$h(t|x, m, \beta) = m t^{m-1} \exp(x'\beta), \quad (2)$$

234 where x represents the independent variables and β represents the
235 parameters. The hazard ratio increases over time if $m > 1$, while it decreases
236 monotonically if $m < 1$. The hazard rate is independent of time if $m = 1$.

237 To avoid the case in which the Weibull distribution does not provide a
238 proper fit, we introduce a semiparametric model, called the Cox proportional
239 hazard model. Instead of assuming the distribution of the data, the Cox
240 model assumes that the hazard ratio is constant over time:

$$h(t|x, \beta) = h_0(t) \exp(x'\beta), \quad (3)$$

241 where $h_0(t)$ is the baseline hazard. Note that as long as the proportional
242 hazard assumption is held, there is no need to know the actual distribution

243 shape of $h_0(t)$.¹³

244 In the actual estimation, we extend the normal survival analysis approach
245 in the following two aspects. First, we include time-varying covariates,
246 while most survival analyses are based upon time-invariant covariates, such
247 as gender. It is problematic to include time-varying variables because
248 this approach usually destroys the exogeneity of covariates (Cameron and
249 Trivedi, 2005). For instance, the unemployment period depends upon the
250 job search strategy, but the job search strategy can be affected by the length
251 of unemployment, while a variation such as seasonal cycle would have no
252 feedback effect similar to this. Nevertheless, we believe our time-varying
253 covariates are closer to the latter example and are sufficiently exogenous to
254 use in the estimation. Second, as floods can be observed repeatedly, we apply
255 a survival analysis of repeated events. Several methods can be utilized to
256 incorporate recurrent events, but we adopt an Anderson-Gill-type recurrent
257 event survival analysis.¹⁴

258 4 Results

259 In this section, we first show the overall results of how different types of
260 forest contribute to mitigating flood occurrence using the Cox and Weibull
261 models. We then conduct additional analyses by dividing the samples in

¹³The details of the model selection can be found in Cameron and Trivedi (2005).

¹⁴For more details, see Amorim and Cai (2015).

262 consideration of possible biases.

263 **4.1** *Effects of forest cover on flood occurrence*

264 The results of the survival analysis are presented in Table 2.¹⁵ We first
265 show estimates for the Cox model. In Column 1, we explore the relationship
266 between flood occurrence and each type of forest without controlling for
267 regional demographic characteristics or precipitation levels. We then include
268 these regional characteristics in the model in Column 2.¹⁶ Finally, we include
269 precipitation variables in the estimation model, as shown in Columns 3 and 4.
270 Column 3 includes the annual average precipitation, while Column 4 includes
271 the maximum daily precipitation. Columns 5 and 6 report the estimation
272 results using models with the Weibull distribution corresponding to Columns
273 3 and 4, respectively.

274 The above results are further confirmed by estimating the parametric
275 model with the assumption of a Weibull distribution. Columns 5 and 6 show
276 the corresponding results, suggesting that broadleaf forest and mixed forest
277 play roles in mitigating the frequency of floods. Comparing the coefficients of

¹⁵We also conducted similar analyses with the total forest area as an independent variable. Results similar to those of our main analyses (Table 2) were obtained. All specifications included subdistrict-fixed effects, which captured unobserved regional characteristics such as distance to the nearest river. To focus on our main objective (i.e., the effects of different forest types on flood occurrence), these results are shown in Table S2 of supplementary material.

¹⁶The regressions of models other than that in Column 1 of Table 2 include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land, in the subdistrict.

278 broadleaf forest and mixed forest, those of broadleaf forest were larger than
279 those of mixed forest; this finding reiterates that broadleaf forest is more
280 effective than other types of forest at mitigating the frequency of floods.

281 **4.2** *Selection bias*

282 Our survival analyses suggest that an increase in forest area has an effect on
283 flood mitigation (Table S2 of supplementary material), particularly increases
284 in the areas of broadleaf forest and mixed forest (Table 2). However, since a
285 gain in forest cover might not occur randomly, there is a possibility that our
286 results suffer from a sample selection bias. For example, there is a possibility
287 that gains in forest cover occurred only in subdistricts where the potential
288 flood risk is low. Therefore, we test for biases by restricting the sample to
289 areas that have a potential flood occurrence risk. Here, we apply only the
290 Cox model, as the Weibull model shows similar results.

291 Table 3 shows the results. The test sample is composed of 107 subdistricts
292 that experienced at least one flood during the study period. The coefficients
293 for broadleaf and mixed forest were negative and statistically significant,
294 while those for coniferous forest were not significant. These results support
295 our findings in Table 2 that broadleaf forest and mixed forest have the
296 potential to mitigate the occurrence of floods, and the broadleaf forest
297 coefficients are similarly larger than the mixed forest coefficients.

298 **4.3** *Heterogeneous effects*

299 Since the probability of flood occurrence increases in response to
300 precipitation, there is a possibility that afforestation policies target
301 high-precipitation areas for the planting of trees. In the case that the
302 estimates suffer from unobserved bias, we test for such bias by dividing the
303 subdistricts based on precipitation. We define high- and low-precipitation
304 areas based on maximum daily rainfall above or below a precipitation
305 threshold of 77 mm in a year. In other words, subdistricts that experienced
306 daily rainfall above 77 mm (sample mean) are defined as high-precipitation
307 areas. The explanatory variables are the same as those in our main analyses.
308 Here, we apply only the Cox model, as the Weibull model shows similar
309 results.

310 Table 4 shows the results for high-precipitation areas in Columns 1 and 2
311 and low-precipitation areas in Columns 3 and 4. The coefficients of broadleaf
312 and mixed forests remained negative and statistically significant in every
313 specification, suggesting that the hydrological effects of forests elucidated
314 above are robust.

315 Similarly, there is a possibility that the flood mitigation effects are
316 different depending on the climate. To test the heterogeneity effects among
317 climates, we estimated the models by dividing the samples into two climate

318 zones based on Li et al. (2013)'s definitions: tropical and monsoon areas
319 and temperate and plateau areas (Figure S1 of supplementary material).¹⁷
320 Columns 5–8 of Table 4 show the results.¹⁸ The coefficients of *mixed forest*
321 remain negative and statistically significant in every area. However, in
322 temperate and plateau areas, the coefficients of *broadleaf forest* are negative
323 but statistically insignificant. This finding suggests that the flood mitigation
324 effects depend on the tree species and ecological characteristics.

325 4.4 *Different levels of severity*

326 In addition, there is a possibility that the tree cover effects on flood mitigation
327 are heterogeneous depending on the intensity of floods because floods occur
328 with multivariate processes. In fact, European Union (2007) emphasizes
329 that a flood management plan should be based on information such as the
330 potential size of the area affected and the depth and velocity of water because
331 they are not independent. Using copula theory, Salvadori et al. (2016) showed
332 the importance of the multivariate flood process in general, while Yin et al.
333 (2018) assessed the implications of climate change in the Ganjiang River
334 basin in China.

¹⁷Several areas are categorized as both monsoon and temperate. Our estimations include these mixed areas in both monsoon and temperate models. This approach has the advantage that our estimations would be more efficient in terms of sample size and degree of freedom.

¹⁸As the Cox model failed to achieve a convergence of the likelihood function, we apply the Weibull models.

335 We test for heterogeneity by applying the estimations to higher and lower
336 intensities of flood events, which correspond to the severity classes reported
337 in the flood dataset (Brakenridge, 2012). The severity of flood events was
338 classified based on the flood recurrence interval: Class 1 includes large
339 floods with reported intervals for one or two decades, and Class 2 includes
340 extreme flood events with reported intervals greater than 100 years. The
341 dependent variable takes the value of one if the flood is categorized as Class
342 2 for high-intensity estimation and Class 1 for low-intensity estimation. The
343 explanatory variables are the same as those in our main analyses presented
344 in Subsection 4.1. Similar to the estimations in Columns 5–8 of Table 4, the
345 Cox model failed to achieve a convergence of the likelihood function; thus,
346 we apply the Weibull model.

347 Table 5 shows the results for high-intensity flood events in Columns 1
348 and 2 and low-intensity flood events in Columns 3 and 4. The coefficients
349 of the broadleaf and mixed forest had a significant negative impact on flood
350 frequency. This finding suggests that the tree cover has mitigation effects on
351 flood frequency, regardless of the flood intensity level.

352 **5 Discussion**

353 Our results are consistent with findings from previous literature on the flood
354 mitigation effects of forest cover (Bradshaw et al., 2007; Bhattacharjee and

355 Behera, 2017, 2018). In addition, our results indicate that the effects on flood
356 occurrences are different depending on the type of tree cover. Broadleaf and
357 mixed forests have mitigation effects, while coniferous forest does not. This
358 finding indicates that increases in the areas of broadleaf and mixed forest
359 have the potential to mitigate the frequency of floods. Furthermore, the
360 absolute values of the coefficients for broadleaf forest were slightly larger than
361 those for mixed forest, suggesting that broadleaf forest is more effective than
362 mixed forest at mitigating flood occurrence. However, increases in the area of
363 coniferous forest are not associated with the mitigation of flood occurrence.
364 Coniferous trees tend to have high market value due to their demand as home
365 building materials. There may be an incentive to plant coniferous trees rather
366 than broadleaf trees at the time of afforestation, as they have higher value
367 when logging after a long time. This study shows that if policy makers make
368 such decisions, they rely too much on trees.

369 Figures 5 helps clarify the effect of each forest type on flood occurrence.
370 These figures illustrate the difference in forest effects between the areas with
371 increasing and decreasing forest cover by forest type based on Nelson-Aalen
372 cumulative hazard estimates. The results indicate that the probability of
373 flood occurrence decreased in areas with increasing broadleaf and mixed
374 forest cover, while this tendency was not observed for coniferous forest.
375 These results are consistent with the findings in the field of forestry science,

376 indicating that broadleaf forest contributes to the mitigation of underground
377 water flow (Komatsu et al., 2007).

378 Other things being equal, the net precipitation (sum of throughfall and
379 stemflow) through a forest is defined by gross precipitation minus total
380 interception loss, which is the sum of canopy interception loss and litter
381 interception loss. When the net precipitation per time reaching the ground
382 exceeds a threshold, a flood occurs (Poorter, 2004). Broadleaf trees usually
383 have more complex shapes and more leaves than coniferous trees. This
384 characteristic enables broadleaf trees to capture more rain and reduce the
385 peak level of net precipitation per time. Precipitation spending more
386 time on leaves and stems increases evapotranspiration as well (Sato, 2007).
387 Combining these two effects, broadleaf forests can reduce the possibility
388 of exceeding the threshold. Broadleaf trees gather precipitation through
389 stemflow, while coniferous trees tend to spread rainfall into relatively broader
390 areas (Kume, 2007). Since soil near a tree is drier due to the consumption of
391 water by the root of the tree, it helps to prevent too much runoff. In addition,
392 changes in forest cover alter not only storm runoff but also base flow (mainly
393 groundwater flow). Yin et al. (2018) discussed that deforestation can increase
394 storm runoff but reduce base flow because the water-holding capacity of the
395 soil decreases when the quality of the forest is degraded. Usually, broadleaf
396 trees generate richer soil with more litter. This characteristic might be

397 another advantage of broadleaf forest.

398 Columns 3 and 4 in Table 2 show the estimated results with the logarithms
399 of the annual average and maximum daily precipitation, respectively, as
400 the explanatory variables. The coefficients of precipitation indicate positive
401 effects on the flood frequency. These results are intuitively reasonable and
402 similar to the conclusions of previous analyses (see Section 2). Furthermore,
403 the coefficients of *GRP* were significantly negative for all the models, meaning
404 that increasing the economic level of a subdistrict has a flood mitigation
405 effect.

406 Overall, our findings remain significant across various model
407 specifications. Specifically, we confirmed that broadleaf trees and mixed-tree
408 forests have effects on flood mitigation, regardless of the precipitation level,
409 climate zones, and flood intensity. This finding suggests that the flood
410 mitigation effects of forests are not particular to certain regions.

411 **6 Conclusion**

412 In this study, we examined the hydrological effects of forests on the mitigation
413 of floods in China, focusing particularly on the effects of different forest types,
414 by applying satellite data to forest and flood data. This study contributes to
415 the literature by estimating how flood prevention effects differ by forest type
416 by applying rigorous survival analysis using samples from the whole country

417 of China. We found that, in accordance with recent hydrological and forestry
418 research, forests moderated the occurrence of floods. We then evaluated the
419 effects by dividing the forest areas by type and found that broadleaf forest
420 and mixed forest contributed to flood prevention, while coniferous forest did
421 not.

422 These results pose important policy implications for policymakers
423 considering flood mitigation by promoting afforestation, which has recently
424 received attention as Eco-DRR. While coniferous forests might not help
425 prevent flooding, coniferous trees tend to be preferred in afforestation policy,
426 as coniferous trees have economic value as wood resources for construction.
427 For example, in the GGP, coniferous trees such as Chinese fir and Masson pine
428 have been preferred (Zhou et al., 2007; Delang and Yuan, 2016). However,
429 in terms of flood prevention, coniferous forests are not effective.

430 In addition, it is worth noting that forests have the potential to mitigate
431 floods over broad areas by leveraging the functions of trees. For example,
432 trees could moderate the yield of water in areas by capturing and recycling
433 precipitation. Hence, considering the effects of forests as Eco-DRR solutions
434 during conventional flood mitigation efforts, such as the construction of
435 levees and dams, might be effective for flood management. These policy
436 implications are applicable not only to China but also to other countries,
437 as the mechanism of flood prevention by forest type can be applied to any

438 country.

439 Finally, several limitations of this study should be mentioned. First,
440 several landscape variations and subdistrict-level variables to control for
441 flood occurrences were excluded from our estimates due to data limitations.
442 Although our time-invariant fixed effects approach captured unobserved
443 regional characteristics such as the distance to the nearest river, there was
444 a possibility of bias due to other omitted variables. For example, we could
445 not include regional investments in flood mitigation, such as the construction
446 of levees and dams, because of the limited availability of data. Therefore,
447 we cannot fully rule out the possibility of bias from unobserved explanatory
448 characteristics on the mitigation of flood occurrence.

449 Second, while this study ascertained the hydrological effects of some
450 forest types, we cannot clearly determine the mechanism underlying the
451 mitigation of flood occurrence. As we discussed in Section 2, how forests
452 mitigate flooding is complex and broadly debated in the fields of forestry
453 science and hydrology. Further studies should attempt to address these
454 issues to promote flood prevention by considering the functions of forests.
455 Nevertheless, although these topics constitute areas of improvement, our
456 study confirms that flood mitigation effects differ by forest type and that
457 broadleaf and mixed forest types are particularly effective; moreover, these
458 findings are robust to our various specifications.

459 Third, we cannot examine the detailed effects of different tree species
460 and vegetation characteristics. The forest cover data we employed include
461 broadleaf, coniferous, and mixed forest. Although there are a variety of
462 tree species and ecological characteristics depending on climate properties,
463 information on detailed tree species is not available. Future studies should
464 attempt to address these issues.

⁴⁶⁵ **Acknowledgment**

⁴⁶⁶ This work was supported by JSPS KAKENHI Grant numbers 19K12467,
⁴⁶⁷ 19H04340, and 18K11754 and National Institute for the Humanities (NIHU
⁴⁶⁸ Transdisciplinary Project “NIHU Area Studies Project for Northeast Asia”).

469 **References**

470 Albert, C., Schröter, B., Haase, D., Brillinger, M., Henze, J., Herrmann, S.,
471 Gottwald, S., Guerrero, P., Nicolas, C., Matzdorf, B., 2019. Addressing
472 societal challenges through nature-based solutions: How can landscape
473 planning and governance research contribute? *Landscape and Urban*
474 *Planning* 182, 12–21.

475 Amorim, L.D., Cai, J., 2015. Modelling recurrent events: a tutorial for
476 analysis in epidemiology. *International Journal of Epidemiology* 44,
477 324–333.

478 An, W., Li, Z., Wang, S., Wu, X., Lu, Y., Liu, G., Fu, B., 2017. Exploring the
479 effects of the “grain for green” program on the differences in soil water in
480 the semi-arid loess plateau of china. *Ecological Engineering* 107, 144–151.

481 Andréassian, V., 2004. Waters and forests: from historical controversy to
482 scientific debate. *Journal of Hydrology* 291, 1–27.

483 Bhattacharjee, K., Behera, B., 2017. Forest cover change and flood hazards
484 in india. *Land Use Policy* 67, 436–448.

485 Bhattacharjee, K., Behera, B., 2018. Does forest cover help prevent flood
486 damage? Empirical evidence from India. *Global Environmental Change*
487 53, 78–89.

488 Bosch, J.M., Hewlett, J., 1982. A review of catchment experiments
489 to determine the effect of vegetation changes on water yield and
490 evapotranspiration. *Journal of Hydrology* 55, 3–23.

491 Bradshaw, C.J.A., Sodhi, N.S., Peh, S.H.K., Brook, B.W., 2007. Global
492 evidence that deforestation amplifies flood risk and severity in the
493 developing world. *Global Change Biology* 13, 2379–2395.

494 Brakenridge, G., 2012. Global active archive of large flood events,
495 Dartmouth Flood Observatory, University of Colorado, available at:
496 <http://floodobservatory.colorado.edu/Archives/index.html> (last access: 8
497 January 2020).

498 Brown, A.E., Zhang, L., McMahon, T.A., Western, A.W., Vertessy, R.A.,
499 2005. A review of paired catchment studies for determining changes in
500 water yield resulting from alterations in vegetation. *Journal of Hydrology*
501 310, 28–61.

502 Bruijnzeel, L.A., 2004. Hydrological functions of tropical forests: Not seeing
503 the soil for the trees? *Agriculture, Ecosystems and Environment* 104,
504 185–228.

505 Calliari, E., Staccione, A., Mysiak, J., 2019. An assessment framework for

506 climate-proof nature-based solutions. *Science of the Total Environment*
507 656, 691–700.

508 CIFOR, 2005. *Forests and floods: drowning in fiction or thriving on facts?*
509 SE - Forest Perspectives. Volume 2. Center for International Forestry
510 Research and Food and Agriculture Organization of the United Nations.

511 Costanza, R., de Groot, R., Sutton, P., Van der Ploeg, S., Anderson, S.J.,
512 Kubiszewski, I., Farber, S., Turner, R.K., 2014. Changes in the global
513 value of ecosystem services. *Global Environmental Change* 26, 152–158.

514 CRED, 2019. *Natural Disasters 2018*. Centre for Research on the
515 Epidemiology of Disasters.

516 Delang, C.O., Yuan, Z., 2016. *China's Grain for Green Program*. Springer.

517 Dorst, H., van der Jagt, S., Raven, R., Runhaar, H., 2019. Urban greening
518 through nature-based solutions—Key characteristics of an emerging
519 concept. *Sustainable Cities and Society* 49, 101620.

520 Ellison, D., Morris, C.E., Locatelli, B., Sheil, D., Cohen, J., Murdiyarso,
521 D., Gutierrez, V., Van Noordwijk, M., Creed, I.F., Pokorny, J., et al.,
522 2017. Trees, forests and water: Cool insights for a hot world. *Global*
523 *Environmental Change* 43, 51–61.

524 European Union, 2007. Directive 2007/60/ec of the european parliament and
525 of the council of 23 october 2007 on the assessment and management of
526 flood risks (text with eea relevance) .

527 Fan, M., Xiao, Y.t., 2020. Impacts of the grain for green program on
528 the spatial pattern of land uses and ecosystem services in mountainous
529 settlements in southwest china. *Global Ecology and Conservation* 21,
530 e00806.

531 FAO, 2018. The state of the world's forests 2018-Forest pathways to
532 sustainable development. Food and Agriculture Organization of the United
533 Nations, Rome.

534 Ferreira, S., Ghimire, R., 2012. Forest cover, socioeconomics, and reported
535 flood frequency in developing countries. *Water Resources Research* 48.

536 Ferreira, S., Hamilton, K., Vincent, J.R., 2013. Does development reduce
537 fatalities from natural disasters? New evidence for floods. *Environment
538 and Development Economics* 18, 649–679.

539 Filoso, S., Bezerra, M.O., Weiss, K.C., Palmer, M.A., 2017. Impacts of forest
540 restoration on water yield: a systematic review. *PloS one* 12, e0183210.

541 Gentry, A., Lopez-Parodi, J., 1980. Deforestation and increased flooding of
542 the upper amazon. *Science* 210, 1354–1356.

543 Güneralp, B., Güneralp, İ., Liu, Y., 2015. Changing global patterns of urban
544 exposure to flood and drought hazards. *Global Environmental Change* 31,
545 217–225.

546 Hewlett, J.D., 1982. *Principles of forest hydrology*. University of Georgia
547 Press.

548 Ickowitz, A., Powell, B., Salim, M.A., Sunderland, T.C., 2014. Dietary
549 quality and tree cover in africa. *Global Environmental Change* 24, 287–294.

550 Kahn, M.E., 2005. The death toll from natural disasters: the role of income,
551 geography, and institutions. *Review of Economics and Statistics* 87,
552 271–284.

553 Keesstra, S., Nunes, J., Novara, A., Finger, D., Avelar, D., Kalantari,
554 Z., Cerdà, A., 2018. The superior effect of nature based solutions in
555 land management for enhancing ecosystem services. *Science of the Total*
556 *Environment* 610, 997–1009.

557 Kellenberg, D.K., Mobarak, A.M., 2008. Does rising income increase or
558 decrease damage risk from natural disasters? *Journal of Urban Economics*
559 63, 788–802.

560 Komatsu, H., Tanaka, N., Kume, T., 2007. Do coniferous forests evaporate

561 more water than broad-leaved forests in japan? *Journal of Hydrology* 336,
562 361–375.

563 Kume, A., 2007. Importance of root systems in forest hydrological processes.
564 Morikita Publishing.

565 Li, G., Sun, S., Han, J., Yan, J., Liu, W., Wei, Y., Lu, N., Sun, Y.,
566 2019. Impacts of Chinese Grain for Green Program and climate change
567 on vegetation in the loess plateau during 1982–2015. *Science of the Total*
568 *Environment* 660, 177–187.

569 Li, X.X., Wang, L.X., Zhang, H., Du, X., Jiang, S.W., Shen, T., Zhang, Y.P.,
570 Zeng, G., 2013. Seasonal variations in notification of active tuberculosis
571 cases in china, 2005–2012. *PLoS One* 8.

572 Lu, Q., Xu, B., Liang, F., Gao, Z., Ning, J., 2013. Influences of the
573 Grain-for-Green Project on grain security in southern china. *Ecological*
574 *Indicators* 34, 616–622.

575 MEA, 2005. *Ecosystems and Human Well-being: Policy Responses*, Volume
576 3. Millennium Ecosystem Assessment, Washington, D.C.

577 Najibi, N., Devineni, N., 2018. Recent trends in the frequency and duration
578 of global floods. *Earth System Dynamics* 9, 757–783.

579 Noy, I., 2009. The macroeconomic consequences of disasters. *Journal of*
580 *Development Economics* 88, 221–231.

581 Ogden, F.L., Crouch, T.D., Stallard, R.F., Hall, J.S., 2013. Effect of land
582 cover and use on dry season river runoff, runoff efficiency, and peak storm
583 runoff in the seasonal tropics of central panama. *Water Resources Research*
584 49, 8443–8462.

585 Peng, J., Hu, X., Wang, X., Meersmans, J., Liu, Y., Qiu, S., 2019. Simulating
586 the impact of Grain-for-Green Programme on ecosystem services trade-offs
587 in Northwestern Yunnan, China. *Ecosystem Services* 39, 100998.

588 Poorter, H., 2004. Larcher, w. physiological plant ecology. *Annals of botany*
589 93, 616.

590 Qian, C., Shao, L., Hou, X., Zhang, B., Chen, W., Xia, X., 2019.
591 Detection and attribution of vegetation greening trend across distinct local
592 landscapes under China’s Grain to Green Program: a case study in Shaanxi
593 Province. *CATENA* 183, 104182.

594 Rodríguez, L.G., Hogarth, N.J., Zhou, W., Xie, C., Zhang, K., Putzel, L.,
595 2016. China’s conversion of cropland to forest program: a systematic
596 review of the environmental and socioeconomic effects. *Environmental*
597 *Evidence* 5, 21.

- 598 Salvadori, G., Durante, F., De Michele, C., Bernardi, M., Petrella, L., 2016.
599 A multivariate copula-based framework for dealing with hazard scenarios
600 and failure probabilities. *Water Resources Research* 52, 3701–3721.
- 601 Sant’Anna, A.A., 2018. Not so natural: unequal effects of public policies on
602 the occurrence of disasters. *Ecological Economics* 152, 273–281.
- 603 Sato, Y., 2007. *Rainfall Inception: How to estimate the amount of water*
604 *reaching to the forest soil.* Morikita Publishing.
- 605 Song, X., Peng, C., Zhou, G., Jiang, H., Wang, W., 2015. Chinese Grain
606 for Green Program led to highly increased soil organic carbon levels: a
607 meta-analysis. *Scientific Reports* 4, 4460.
- 608 Sulla-Menashe, D., Gray, J.M., Abercrombie, S.P., Friedl, M.A., 2019.
609 Hierarchical mapping of annual global land cover 2001 to present: The
610 MODIS Collection 6 Land Cover product. *Remote Sensing of Environment*
611 222, 183 – 194.
- 612 Swank, W.T., Douglass, J.E., 1974. Streamflow greatly reduced by converting
613 deciduous hardwood stands to pine. *Science* 185, 857–859.
- 614 Tan-Soo, J.S., Adnan, N., Ahmad, I., Pattanayak, S.K., Vincent, J.R.,
615 2016. Econometric evidence on forest ecosystem services: Deforestation

616 and flooding in Malaysia. *Environmental and Resource Economics* 63,
617 25–44.

618 Van Dijk, A.I., van Noordwijk, M., Calder, I.R., Bruijnzeel, S.L., Schellekens,
619 J., Chappell, N.A., 2009. Forest–flood relation still tenuous–comment on
620 ‘Global evidence that deforestation amplifies flood risk and severity in the
621 developing world’ by C. J. A. Bradshaw, N. S. Sodi, K. S.-H. Peh and
622 B.W. Brook. *Global Change Biology* 15, 110–115.

623 Van Westen, C.J., Castellanos, E., Kuriakose, S.L., 2008. Spatial data for
624 landslide susceptibility, hazards and vulnerability assessment: an overview.
625 *Engineering Geology* 102, 112–131.

626 Wallemacq, P., House, R., 2018. UNISDR and CRED report: Economic
627 losses, poverty & disasters (1998–2017). Centre for Research on the
628 Epidemiology of Disasters.

629 Wang, Y., Zhao, J., Fu, J., Wei, W., 2019. Effects of the Grain for Green
630 Program on the water ecosystem services in an arid area of China-using the
631 Shiyang River Basin as an example. *Ecological Indicators* 104, 659–668.

632 Wells, J.A., Wilson, K.A., Abram, N.K., Nunn, M., Gaveau, D.L.A.,
633 Runting, R.K., Tarniati, N., Mengersen, K.L., Meijaard, E., 2016. Rising

- 634 floodwaters: mapping impacts and perceptions of flooding in Indonesian
635 Borneo. *Environmental Research Letters* 11, 064016.
- 636 Wu, X., Wang, S., Fu, B., Feng, X., Chen, Y., 2019. Socio-ecological changes
637 on the loess plateau of china after grain to green program. *Science of the*
638 *Total Environment* 678, 565–573.
- 639 Xu, Z., Wei, H., Fan, W., Wang, X., Huang, B., Lu, N., Ren, J., Dong, X.,
640 2018. Energy modeling simulation of changes in ecosystem services before
641 and after the implementation of a Grain-for-Green program on the Loess
642 Plateau—a case study of the Zhifanggou valley in Ansai County, Shaanxi
643 Province, China. *Ecosystem Services* 31, 32–43.
- 644 Yamamoto, Y., Shigetomi, Y., Ishimura, Y., Hattori, M., 2019. Forest
645 change and agricultural productivity: Evidence from indonesia. *World*
646 *Development* 114, 196–207.
- 647 Ye, L., Fang, L., Shi, Z., Deng, L., Tan, W., 2019. Spatio-temporal dynamics
648 of soil moisture driven by 'Grain for Green' program on the Loess Plateau,
649 China. *Agriculture, Ecosystems and Environment* 269, 204–214.
- 650 Yin, H., Li, C., 2001. Human impact on floods and flood disasters on the
651 Yangtze River. *Geomorphology* 41, 105–109.
- 652 Yin, J., Guo, S., He, S., Guo, J., Hong, X., Liu, Z., 2018. A copula-based

653 analysis of projected climate changes to bivariate flood quantiles. Journal
654 of Hydrology 566, 23–42.

655 Zhou, S., Yin, Y., Xu, W., Ji, Z., Caldwell, I., Ren, J., 2007. The costs and
656 benefits of reforestation in liping county, guizhou province, china. Journal
657 of Environmental Management 85, 722–735.

658 Zong, Y., Chen, X., 2000. The 1998 flood on the Yangtze, China. Natural
659 Hazards 22, 165–184.

Table 1: Summary statistics

| Variable | Mean | Std. Dev. | Min. | Max. |
|--|----------|-----------|-------|----------|
| <i>Flood</i> | 0.028 | 0.165 | 0 | 1 |
| <i>Broadleaf forest</i> (thousand km ²) | 1.337 | 3.495 | 0 | 28.778 |
| <i>Coniferous forest</i> (thousand km ²) | 0.315 | 1.327 | 0 | 23.243 |
| <i>Mixed forest</i> (thousand km ²) | 5.033 | 13.302 | 0 | 155.273 |
| <i>Maximum daily precipitation</i> (mm) | 77.355 | 39.933 | 0.194 | 355.831 |
| <i>Annual average precipitation</i> (mm) | 825.920 | 466.907 | 0.357 | 2731.924 |
| <i>GRP in the subdistrict</i> (CNY 100 million) | 16.069 | 16.511 | 0.139 | 89.705 |
| <i>Population</i> | 5328.138 | 2754.722 | 264 | 11169 |

Note: The number of observations is 5763.

Table 2: Survival analysis on flood occurrence (all samples).

| | Cox model (1) | Cox model (2) | Cox model (3) | Cox model (4) | Weibull model (5) | Weibull model (6) |
|--|--------------------|----------------------|----------------------|----------------------|----------------------|----------------------|
| <i>Broadleaf forest</i> | -0.010* (0.005) | -0.044*** (0.011) | -0.047*** (0.012) | -0.045*** (0.011) | -0.074*** (0.018) | -0.072*** (0.017) |
| <i>Coniferous forest</i> | -0.004 (0.019) | 0.000 (0.017) | -0.003 (0.018) | -0.003 (0.018) | 0.011 (0.022) | 0.008 (0.020) |
| <i>Mixed forest</i> | -0.008* (0.005) | -0.042*** (0.009) | -0.043*** (0.009) | -0.042*** (0.010) | -0.064*** (0.015) | -0.063*** (0.014) |
| <i>GRP (/1000)</i> | | -0.094*** (0.023) | -0.091*** (0.022) | -0.093*** (0.022) | -0.184*** (0.029) | -0.182*** (0.029) |
| <i>Population</i> | | 0.002*** (0.001) | 0.002*** (0.001) | 0.002*** (0.001) | 0.003*** (0.001) | 0.003*** (0.001) |
| $\ln(\text{Annual average precipitation})$ | | | 2.198*** (0.600) | | 1.590*** (0.607) | |
| $\ln(\text{Maximum daily precipitation})$ | | | | 0.562** (0.277) | | 0.670** (0.275) |
| Observations | 5763 | 5763 | 5763 | 5761 | 5761 | 5761 |
| Log-likelihood | -833.673 | -808.943 | -800.318 | -806.847 | 632.233 | 627.937 |
| Wald chi-square | 12568.099 | 72559.806 | 2314.732 | 11428.741 | | |

Note: The dependent variable is flood occurrence.

Standard errors in parentheses are clustered at the subdistrict level.

***, **, and * denote statistical significance at the 1%, 5%, and 10% levels.

All estimates include subdistrict-fixed effects.

All regressions in Columns 2-6 include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land in each subdistrict.

Table 3: Effects restricted to areas that experienced floods during the study period.

| | (1) | (2) |
|--|----------------------|----------------------|
| <i>Broadleaf forest</i> | -2.111*** (0.524) | -2.012*** (0.509) |
| <i>Coniferous forest</i> | -0.125 (0.783) | -0.112 (0.804) |
| <i>Mixed forest</i> | -1.901*** (0.418) | -1.874*** (0.427) |
| <i>GRP (/1000)</i> | -0.091*** (0.022) | -0.093*** (0.022) |
| <i>Population</i> | 0.002*** (0.001) | 0.002*** (0.001) |
| $\ln(\text{Annual average precipitation})$ | 2.198*** (0.602) | |
| $\ln(\text{Maximum daily precipitation})$ | | 0.562** (0.278) |
| Observations | 1819 | 1819 |
| Log-likelihood | -800.318 | -806.847 |
| Wald chi-square | 2299.880 | 11355.400 |

Note: The dependent variable is flood occurrence.

Standard errors in parentheses are clustered at the subdistrict level.

***, **, and * denote statistical significance at the 1%, 5%, and 10% levels.

All estimates include subdistrict-fixed effects.

All regressions include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land in each subdistrict.

Table 4: Heterogeneous effects by dividing areas based on precipitation levels and climate zones.

| | Precipitation levels | | | Climate zones | | | | |
|--|--------------------------|-------------------------|----------------------|----------------------|-----------------------|----------------------|---------------------|---------------------|
| | High-precipitation areas | Low-precipitation areas | | Monsoon and tropical | Temperate and plateau | | | |
| | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| <i>Broadleaf forest</i> | -3.391** (1.620) | -3.436** (1.423) | -2.048** (0.994) | -1.762* (1.028) | -2.637*** (0.776) | -2.673*** (0.747) | -1.132 (1.252) | -1.006 (1.099) |
| <i>Coniferous forest</i> | -1.124 (2.068) | -1.557 (2.078) | 1.629 (1.719) | 1.270 (2.011) | 0.105 (1.048) | -0.124 (1.004) | -0.400 (11.568) | -2.374 (10.642) |
| <i>Mixed forest</i> | -3.320** (1.568) | -3.378** (1.403) | -1.192*** (0.456) | -1.261** (0.541) | -2.406*** (0.726) | -2.503*** (0.712) | -1.223** (0.546) | -1.187** (0.541) |
| <i>GRP (/1000)</i> | -0.127*** (0.033) | -0.125*** (0.032) | -0.063 (0.041) | -0.065 (0.046) | -0.222*** (0.034) | -0.222*** (0.035) | -0.181* (0.109) | -0.177 (0.109) |
| <i>Population</i> | 0.003*** (0.001) | 0.002*** (0.001) | -0.002 (0.002) | -0.002 (0.002) | 0.004*** (0.001) | 0.004*** (0.001) | 0.001 (0.004) | 0.002 (0.004) |
| $\ln(\text{Annual average precipitation})$ | 2.280*** (0.850) | | 3.368*** (1.073) | | 2.041*** (0.548) | | 1.812* (1.077) | |
| $\ln(\text{Maximum daily precipitation})$ | | 1.340*** (0.366) | | 0.906 (0.721) | | 1.106*** (0.278) | | 1.833** (0.776) |
| Observations | 2463 | 2463 | 3300 | 3300 | 4811 | 4810 | 2839 | 2839 |
| Log-likelihood | -376.002 | -376.110 | -252.453 | -257.655 | 552.236 | 551.717 | 158.379 | 161.169 |
| Wald chi-square | 1774.964 | 2483.771 | 2605.169 | 8558.731 | | | | |

Note: The dependent variable is flood occurrence.

Standard errors in parentheses are clustered at the subdistrict level.

***, **, and * denote statistical significance at the 1%, 5%, and 10% levels.

All estimates include subdistrict-fixed effects.

All regressions include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land in each subdistrict.

Estimates use models with the Cox distribution for Column 1–4 and Weibull distribution for Column 5–8.

Table 5: Effects on different levels of severity of floods.

| | Higher intensity | | Lower intensity | |
|--|----------------------|----------------------|----------------------|----------------------|
| | 1 | 2 | 3 | 4 |
| <i>Broadleaf forest</i> | -11.349** (4.957) | -11.513** (5.265) | -4.841*** (1.540) | -4.396*** (1.566) |
| <i>Coniferous forest</i> | -10.739 (8.811) | -11.479 (8.684) | 1.896 (1.609) | 1.571 (1.624) |
| <i>Mixed forest</i> | -11.764** (5.328) | -12.004** (5.727) | -3.251*** (1.030) | -3.173*** (1.038) |
| <i>GRP (/1000)</i> | -0.287** (0.141) | -0.296** (0.136) | -0.343*** (0.099) | -0.342*** (0.102) |
| <i>Population</i> | 0.004 (0.002) | 0.004* (0.002) | 0.004*** (0.001) | 0.004*** (0.001) |
| $\ln(\text{Annual average precipitation})$ | 2.812* (1.567) | | 2.060*** (0.644) | |
| $\ln(\text{Maximum daily precipitation})$ | | 0.508 (0.679) | | 1.121*** (0.395) |
| Observations | 5762 | 5760 | 5762 | 5760 |
| Log-likelihood | 93.457 | 91.552 | 390.643 | 390.412 |

Note: The dependent variables are flood occurrence, with higher severity corresponding to severity Class 2 in columns 1 and 2 and lower severity corresponding to severity Class 1 in columns 3 and 4.

Standard errors in parentheses are clustered at the subdistrict level.

***, **, and * denote statistical significance at the 1%, 5%, and 10% levels.

All estimates include subdistrict-fixed effects.

All regressions include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land in each subdistrict.

All estimates use models with the Weibull distribution.

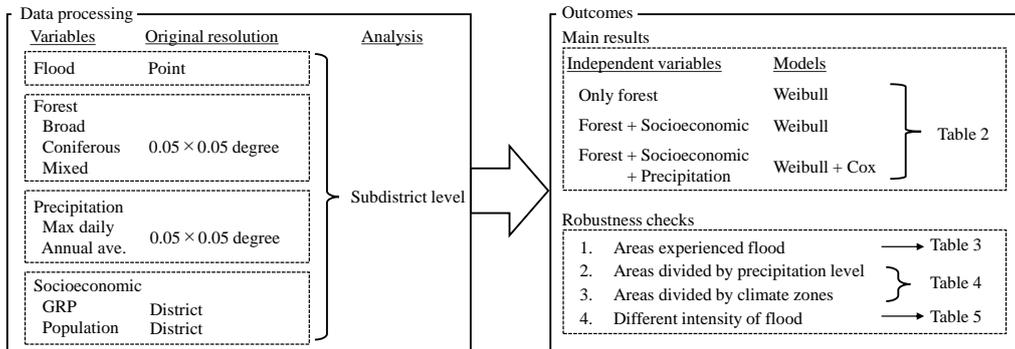


Figure 1: Flowchart of the estimation procedure.

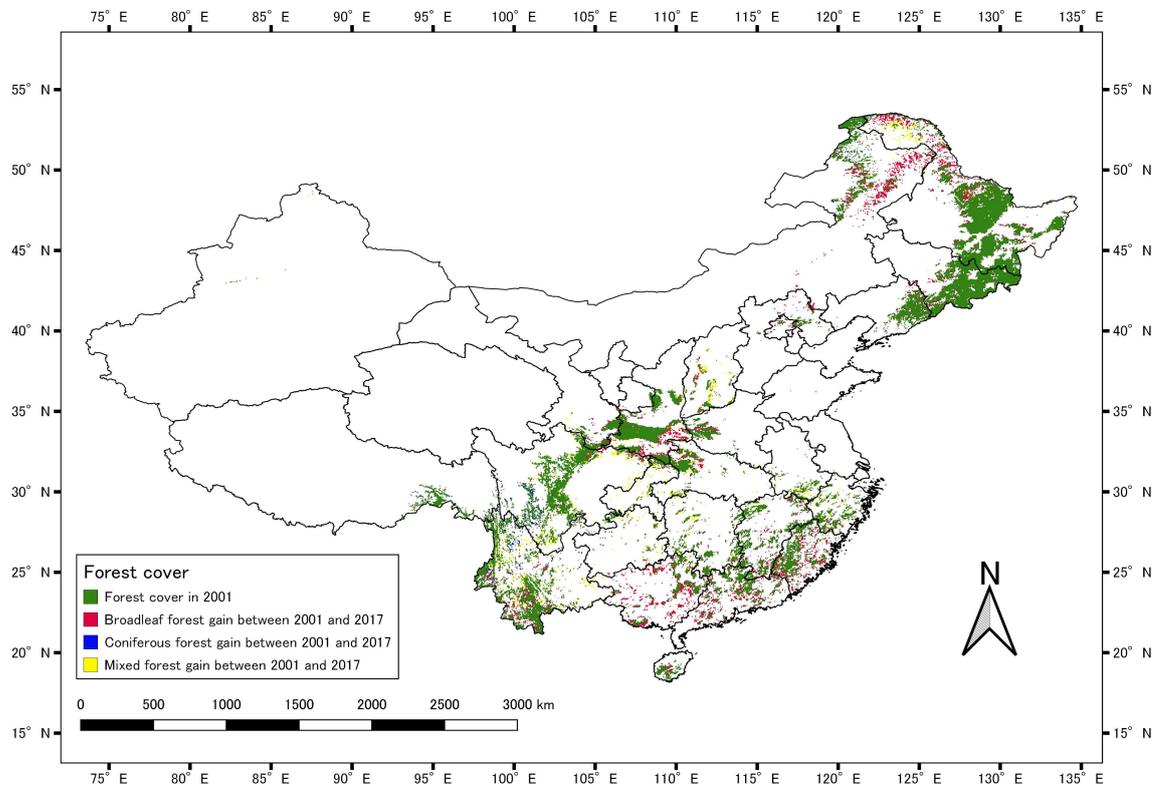


Figure 2: Forest cover and forest gain in China (grid base). *Source: Sulla-Menashe et al. (2019).*

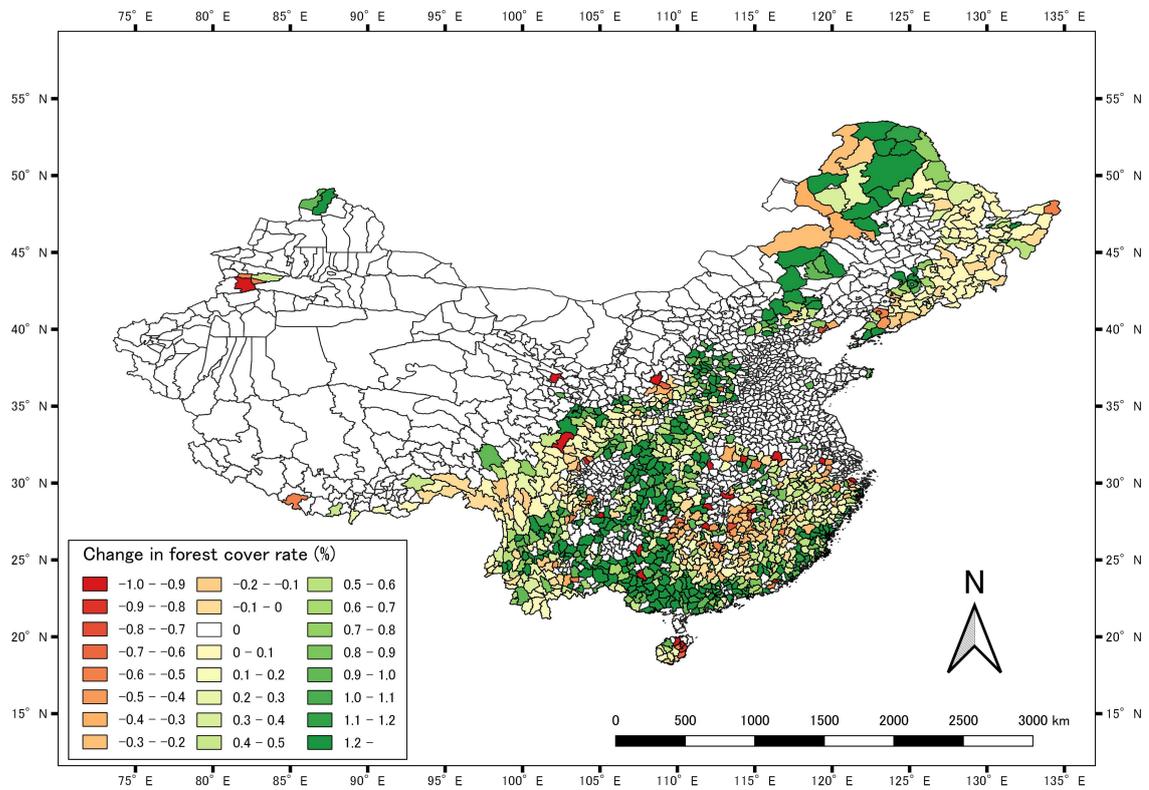


Figure 3: Percentage change in forest cover rate by subdistrict. *Source: Sulla-Menashe et al. (2019).*

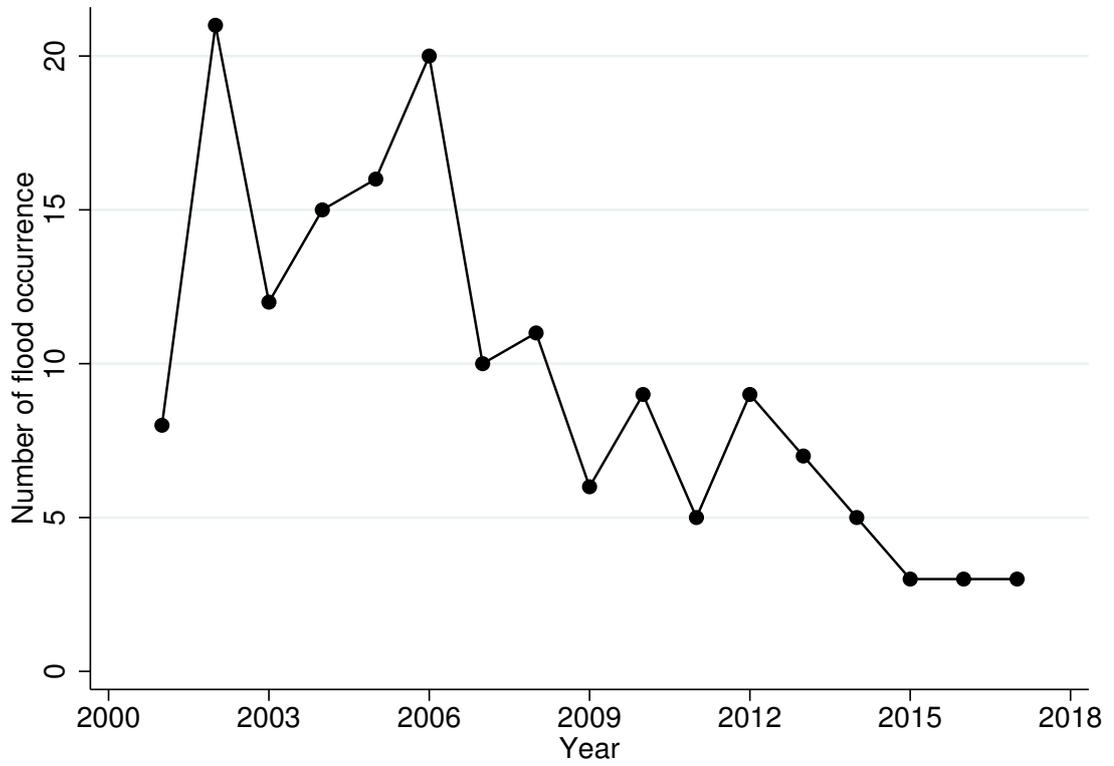


Figure 4: Number of floods that have occurred in China. *Source: Brakenridge (2012).*

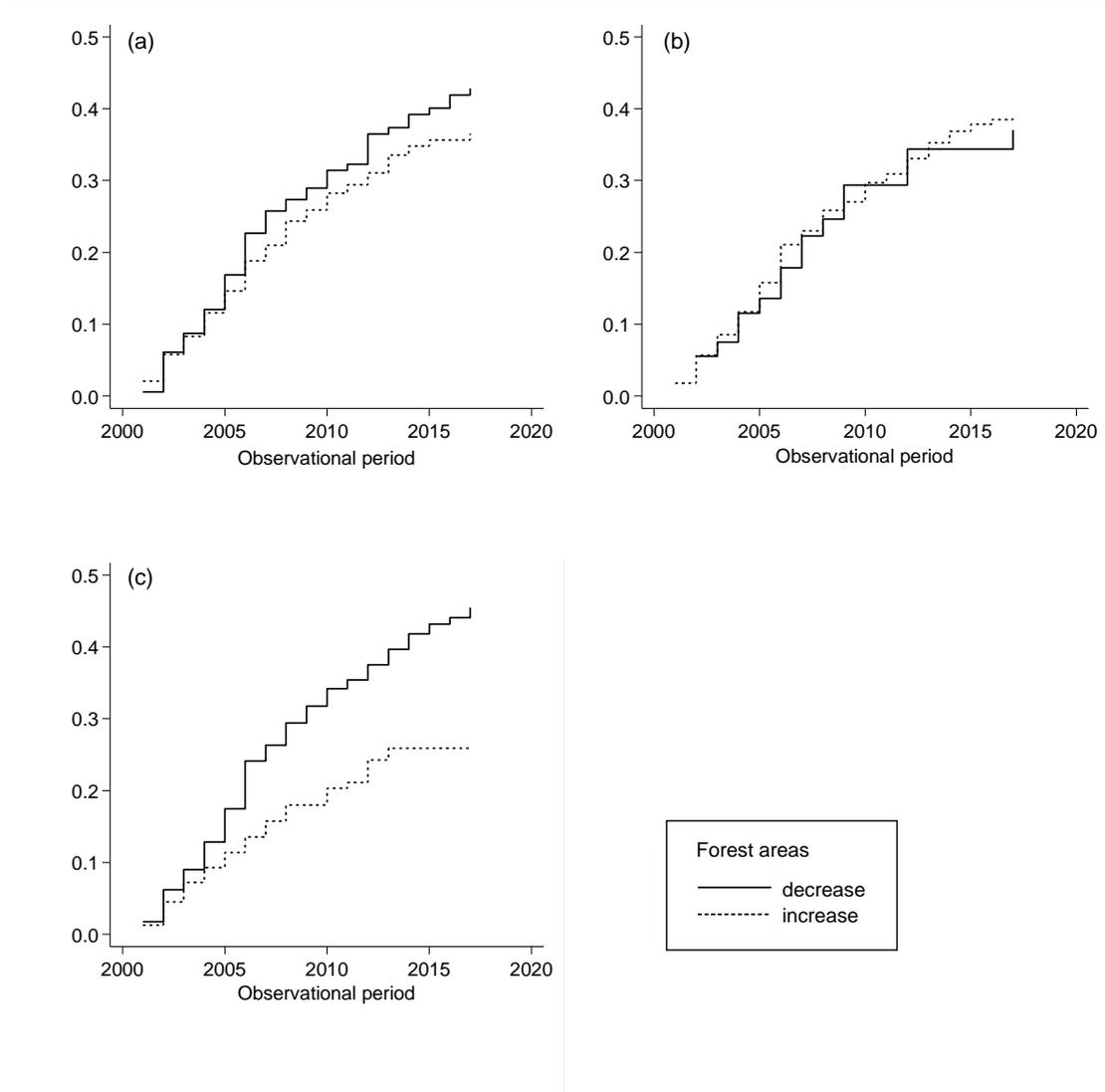


Figure 5: Nelson-Aalen cumulative hazard estimates for broadleaf forest (a), coniferous forest (b), and Mixed forest (c).

660 Supplementary material

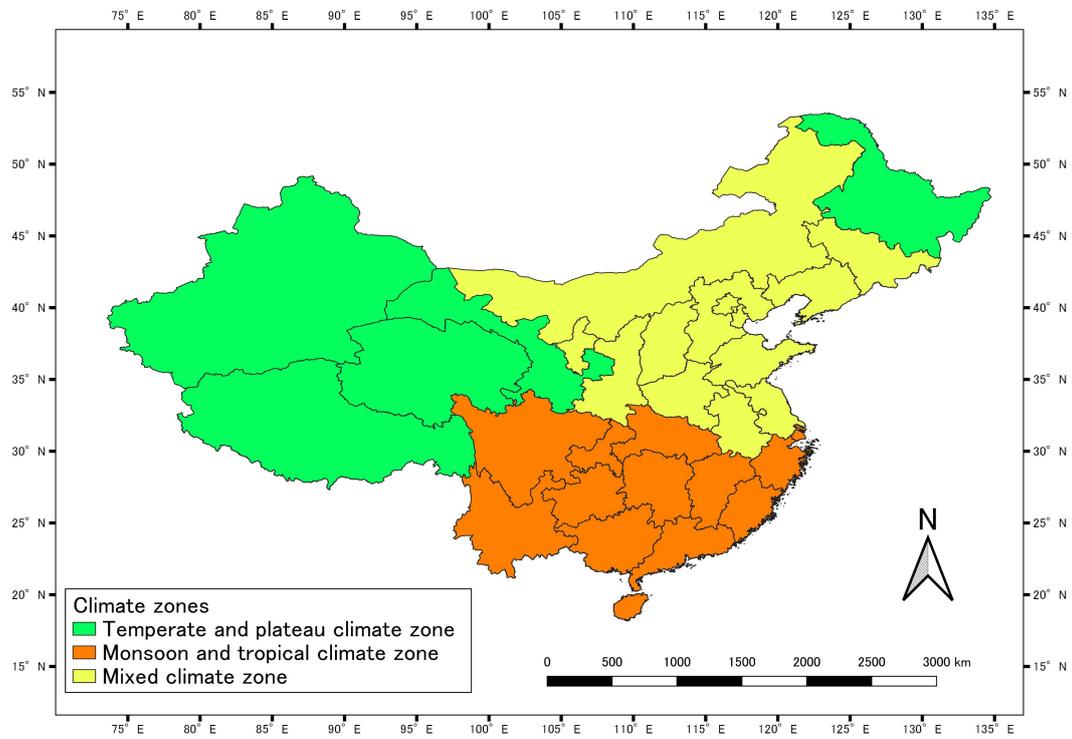


Figure S1: Climate regions in China. *Source: Li et al. (2013)*

Table S1: Variable description and data sources.

| Variable | Description | Original resolution | Data source |
|---|--|----------------------------|---|
| <i>Flood</i> | Recorded flood events | Point data | Brakenridge (2012) |
| <i>Broadleaf forest</i> <i>Coniferous forest</i> <i>Mixed forest</i> <i>Total forest areas</i> | Forest area based on IGBP classification | 0.05×0.05 degrees | Sulla-Menashe et al. (2019) |
| <i>GRP</i> | Gross regional product per thousand CNY | District level | National Bureau of Statistics of China |
| <i>Population</i> | Total population | | |
| <i>Annual average precipitation</i> <i>Maximum daily precipitation</i> | Mean value on geographical approximate grid of subdistrict Maximum value at each subdistrict and year | 0.05×0.05 degrees | Global Unified Precipitation dataset from the Climate Prediction Center |

Table S2: Effect of total forest area on flood occurrence.

| | Cox model (1) | Cox model (2) | Cox model (3) | Cox model (4) | Weibull model (5) | Weibull model (6) |
|---|---------------------|----------------------|----------------------|----------------------|----------------------|----------------------|
| <i>Total forest area</i> | -0.355** (0.165) | -1.454*** (0.286) | -1.564*** (0.299) | -1.586*** (0.306) | -2.059*** (0.607) | -2.064*** (0.536) |
| <i>GRP (/1000)</i> | | -0.092*** (0.022) | -0.090*** (0.022) | -0.091*** (0.022) | -0.180*** (0.027) | -0.177*** (0.028) |
| <i>Population</i> | | 0.002*** (0.001) | 0.002*** (0.001) | 0.002*** (0.001) | 0.003*** (0.001) | 0.003*** (0.001) |
| <i>ln(Annual average precipitation)</i> | | | 2.312*** (0.580) | | 1.774*** (0.502) | |
| <i>ln(Maximum daily precipitation)</i> | | | | 0.655** (0.289) | | 0.728** (0.295) |
| Observations | 5763 | 5763 | 5763 | 5761 | 5763 | 5761 |
| Log-likelihood | -833.665 | -810.951 | -801.411 | -808.182 | 625.035 | 621.650 |
| Wald chi-square | 15137.657 | 18777.887 | 12678.017 | 41728.748 | | |

Note: The dependent variable is flood occurrence.

Standard errors in parentheses are clustered at the subdistrict level.

***, **, and * denote statistical significance at the 1%, 5%, and 10% levels.

All estimates include subdistrict-fixed effects.

All regressions in Columns 2–6 include other land cover areas classified by the IGBP, such as grasslands, croplands, and barren land in each subdistrict.